# Nearshore Habitats as Nursery Grounds for Recreational Important Fishes 

Recreational Fishery Habitat Assessment Study

October 1, 1995 to September 30, 2000

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## FINAL REPORT

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Final Report Job F7<br>NEARSHORE HABITATS AS NURSERY GROUNDS FOR RECREATIONAL IMPORTANT FISHES.<br>Ivan Mateo<br>USVI Division of Fish and<br>Wildlife Room 203 Lagoon Street<br>Complex, St. Croix, VI 00841


#### Abstract

Six protected backreef embayments on St. Croix's northeast and southeast coasts were sampled to determine species composition and juvenile fish abundance, The study sites on the northeast end consisted of Cottongarden Bay, Teague Bay and Yellowcliff Bay. They were sampled monthly from October 1998 to September 1999. The study sites on the southeast end of St. Croix consisted of Turner Hole Bay, Robin Bay, and Great Pond Bay. These were sampled monthly from July 2000 to September 2000. Juvenile reef fish assemblages were sampled using three complementary sampling methods: visual strip transect census, fish traps, and beach seine net. By comparing juvenile fish communities from distinct habitats (patch reef, seagrass, rubble, algal plains, and sand) within the six embayments, significant differences in fish densities and number of species per area were identified. Patch reef habitats had more species per area than any other habitat. Juvenile fish commonly observed in embayments included scarids, labrids and haemulids, The slippery dick, Halichoeres bivittatus, the bucktooth parrotfish, S. radians, the spotted goatfish, Pseudupeneus maculatus, and newly settled Haemulon unknown were among the most abundant species found in embayments. The juvenile phase of certain economically important reef fishes appear to prefer these nearshore embayments.


Key Words: juvenile fish, nursery, and recruitment

## INTRODUCTION

The role of coastal habitats as nursery grounds for coral reef fishes and many invertebrates is well documented (Springer and Mcerlean 1962; Austin 1971; Macnae 1974; Beumer 1978; Rooker and Dennis 1991; and Dennis 1992): Nearshore habitats such as mangrove systems, seagrass meadows, and backreef areas are composed of multi-species communities. The ecological relationships within these communities are wide-ranging and complex. Basically, they provide shelter, refuge, and food for a great number of organisms (Phillips and Menez 1988) including early stages of many organisms (Ogden and Zieman 1977; Shulman 1984 and 1985; Boulon 1992; and Lindeman 1999). The ecological importance of nearshore habitats in backreef areas as nursery areas for recreational and commercial important fishes have been well documented (Ogden and Zieman 1977; Zieman et al. 1982; Robblee and Zieman 1984; Rooker and Dennis 1991; and Nagelkerken et al 2000). Economically important species such as grunts, parrotfishes, some species of snappers and groupers, spiny lobsters, and queen conch have been reported to use these habitats for nursery grounds (Boulon 1986; Stoner and Waite 1990; and Appeldoorn et al 1997).

This report documents results of a study of fish assemblages in nearshore habitats within backreef lagoon areas on the northeast and southeast ends of St. Croix, U.S. Virgin Islands. This information is essential in the management of these areas to sustain and enhance their ecological and fisheries value as nursery grounds.

## Description of Study Sites

The three embayments on the northeast end of St. Croix sampled in this study included: Cottongarden Bay, Teague Bay, and Yellowcliff Bay. These bays are part of the Teague Bay bank-barrier reef system that extends from Pull Point to Lamb Point on the northeast coast of St. Croix (Figure 1). This extensive, nearly continuous reef complex forms a protective barrier against wind and wave energy for numerous shallow backreef embayments. The zonation pattern in these three embayments follows the characteristics of a typical backreef lagoon (see Burke et al 1989). The lagoon is characterized by seagrass and sediment-dwelling organisms dominated by molluscs and echinoderms. The dominant calcareous algae include Penicillus spp. and Halimeda spp. The abrupt transition from the lagoon to the backreef is marked by scattered corals (Montastrea annularis and Porites spp.), several species of gorgonians, and fleshy algae. Lagoon depths range from a maximum of approximately 8 m (west of Teague Point) to 1 m in the backreef at both eastern and western ends of the lagoon.

The three embayments on the southeast end of St. Croix sampled in this study included: Turner Hole Bay, Robin Bay, and Great Pond Bay. These embayments are part of the southeast end bank-barrier reef system that extends from East Point to Vagthus Point (Figure 2). Turner Hole and Robin Bay are similar (Hubbard 1989). They are almost 1000 m long and approximately 300 m wide over most of their lengths. Most of the bottom of the lagoon is covered by seagrasses (Thalassia testudinum and Syringodium
filiforme) occurring in beds of varying density. Halimeda spp., Penicillus spp. and various species of turf like algae are also locally abundant within the seagrass beds. In the deeper portions of the lagoon, areas of sand or thin grass cover are dominated by 10 to 20 cm high sand mounds formed by the burrowing shrimp, Callianassa spp. Other benthic organisms include widely scattered colonies of Porites astreoides, $M$ annularis, and Diploria spp. The third bay, Great Pond Bay, is approximate 2 km long and bounded on its landward side by the baymouth bar and seaward by a continuous coralalgal reef (Bruce et al 1989). The inshore portion of the lagoon (a distance of 30 to 100 m from shore) is covered by the seagrass, T. testudinum, and lesser amounts of S. filiforme. Seaward of that, the majority of the lagoon floor is sand with numerous sand mounds produced by the burrowing shrimp Callianassa spp. Scattered patches of S. filiforme occur along with the algae, Dictyota spp., and Penicillus spp. Scattered coral heads, including M. annularis, Diploria strigosa and Siderastrea spp., also occur in the lagoon.

## METHODS

## Sampling

Sampling Periods - The nearshore nursery habitats in three protected backreef embayments on St. Croix's northeast coast (Cottongarden Bay, Teague Bay and Yellowcliff Bay) were sampled monthly from October 1998 to September 1999. The embayments on the southeast end of St. Croix (Turner Hole Bay, Robin Bay and Great Pond Bay) were sampled monthly from July 2000 to September 2000. This limited sampling period in-the southeast was due to staff shortages and time constraints.

Sampling Grid - For each bay, a $20 \times 20 \mathrm{~m}$ grid pattern was laid over a nautical chart of each embayment. Grid intersecting points were labeled with consecutive numbers. This numbered grid was the basis for selecting transect survey sites and trap sites (see below).

Transect Sites - Once a month during the survey period, 10 sites per embayment were randomly selected based on the numbered grid pattern described above. The sample size (10) for transects was based on a preliminary fish census (see Rogers et al 1994).

Prior to each transect survey, the compass bearing for the transect line was randomly selected. GPS coordinates were not recorded for each transect site. Instead, each location was determined by line of sight with landmarks on the shore. Once at the site, one end of the 50 m transect line tape (marked at 1 cm intervals) was dropped in the water (using a small weight), and the transect line tape was laid by the diver in the direction of the randomly selected compass bearing. A different compass bearing was randomly selected for each of the 10 monthly transect sites in a month. At each transect site, two different surveys were completed: (1) a benthic survey, and (2) a fish census. These are discussed
in detail below.

Any transect site that had previously been surveyed was discarded and another randomly selected site was chosen until 10 previously not sampled sites were selected. Sites selected for fish traps (see below) were excluded from the universe of possible transect sites within each embayment.

Each month, 10 additional sites were randomly selected as new transect sites for each embayment. The compass bearing of each new transect site was randomly selected.

Benthic Surveys - Benthic substrate surveys were conducted at each transect site. The benthic habitat categories selected for this survey included: patch reef, rubble, sand, algal plain, and seagrass (see Adams and Ebersole, in press). These habitat classifications are defined in Appendix A.

Habitat categories were recorded along the 50 m transect line to the nearest cm . The transect line reading (to the nearest cm ) was recorded when habitat category changed. For areas of mixed substrate composition, the habitat category that dominated the particular area was recorded as the habitat category.

The percentage composition of each habitat category (for each transect) was estimated by summing the length of line with a habitat category and dividing by the total length of the transect line ( 50 m ).

Fish Census - At each transect site, a fish census was completed by two divers swimming the 50 m transect line. One diver inspected 2 m to the left of the transect line. The other diver inspected 2 m to the right of the transect line. For the 50 m transect line a total of $200 \mathrm{~m}^{2}$ ( 50 mx 4 m ) was surveyed at each site, Each diver recorded major habitat category type (see Appendix A), fish species present, number of fish present for each species, and size categories of fish observed along the 50 m transect line to the nearest cm .

For most fish species, size categories used here were $<5 \mathrm{~cm}$ total length (TL), 5-10 cm TL, and $>10 \mathrm{~cm}$ TL. For small fish species such as wrasse, grunts, gobies, blennids, cardinal fish and damselfish, an additional size category was added ( $<3 \mathrm{~cm} \mathrm{TL}$ ). For these small fish species, fish $<3 \mathrm{~cm}$ TL were identified as newly settled fish, Since fish $<3 \mathrm{~cm}$ TL were difficult to identify to the species level, identification was limited to the genus level. These are indicated here as species unknown. Data was pooled at the genus level for subsequent analysis.

For each transect site, fish densities within habitats were calculated by summing the number of fish within a habitat and dividing by the total area of that habitat within the transect area ( $50 \mathrm{mx} 4 m \times$ percentage of a habitat from benthic survey).

Fish Traps - Fish traps used in this study were rectangular ( $92 \mathrm{~cm} \times 57 \mathrm{~cm} \times 19 \mathrm{~cm}$ ) and made from vinyl-coated wire with 1.3 cm bar mesh. Each trap had one escape panel ( 15 x 10 cm ). The outside rectangular funnel entrance opening measured 23 cm high $\times 25$ cm wide, The funnel was 45.7 cm deep from the outside rectangular opening to the inside oval entrance, The inside funnel opening was 10.2 cm high by 7.6 cm wide. Each trap was baited with approximately 0.5 lb herring per trap.

For each embayment, ten fish trap sites were randomly selected each month based on the $20 \mathrm{~m} \times 20 \mathrm{~m}$ grid (see above). If a grid number was selected as a transect site (see above), it was omitted and never used as a fish trap site. Each trap was fished for 24 hours. All fish caught were identified and measured (total length) to the nearest millimeter, then released. The habitat where each fish trap was set was not recorded.

Each month, 10 new fish trap sites were randomly selected for each embayment. Once a grid site was used as a trap site, that grid site was omitted and not included in universe for subsequent trap site selection.

Beach Seine - The beach seine net measured $30.5 \mathrm{~m} \times 122 \mathrm{~cm}$ and had weights and floats attached. The net mesh size was 1.3 cm stretch mesh. Two PVC pipes (1.75 meters long each) were fixed vertically at each end of the net.

Two beach seine hauls were conducted monthly in each embayment. In the first month, the two seine-hauls were done on the east side of each embayment. The following month, the two seine-hauls were done on the west side of each embayment. For subsequent months, the seine haul sites alternated from the east side to the west side of the embayment. At each seine-haul site, the two seine-hauls were done about 10 to 20 m apart.

For each seine-haul, one end of the seine net was fixed onshore. The other end of the net was manually pulled out into the water perpendicular to shore. Once the net was fully extended, the seaward-end of the seine net was pulled to shore in an are with the shoreend of the net fixed. The total area of the sweep was about $730 \mathrm{~m}^{2}\left(\pi \times 30.5 \mathrm{~m}^{2} / 4\right)$. All fish caught in the seine net were identified, enumerated, measured (total length) to the nearest millimeter, and released at the point of capture.

## Data Analysis

Visual Census - The distribution of fish density and number of species per area estimates (pooled embayment, transect site and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If fish density and number of species per area estimates failed the normality test even after log
$(x+1)$ and square root transformations, non-parametric statistics were used to analyze the data. Fish density and number of species per area estimates for each embayment were then compared using the non-parametric Kruskall-Wallis One-Way ANOVA on ranks (Sokal and Rohlf 1981). If significant differences were detected, then Dunn's multiple comparison procedure was used to detect differences in mean fish density and number of species per area between embayments.

The Shannon-Weaver Diversity Index H' (Shannon and Weaver 1949) and the evenness index J' (Pielou 1978) were applied to fish density estimates for each habitat and embayment. These two indices were calculated for each northeast St. Croix embayment by month (pooled transect site data), then for each northeast St. Croix embayment by habitat (pooled transect site and month data), Since southeast St. Croix embayments were only surveyed for 3 months, these indices were calculated on fish density data (pooled by transect site and month).
$\mathrm{H}^{\prime}$ and $\mathrm{J}^{\prime}$ estimates (pooled embayment, transect site and monthly data) were then checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If $\mathrm{H}^{\prime}$ and $\mathrm{J}^{\prime}$ estimates failed the normality test even after $\log (\mathrm{x}+1)$ and square root transformations, the Kruskall-Wallis One-way ANOVA was used to compare data. If significant differences were detected then Dunn's multiple comparison procedure was used to detect differences in diversity and evenness indices between habitats.

Differences in densities of fish by fish length (size class) among habitats and embayments were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If fish densities by size class failed the normality test even after $\log (x+1)$ and square root transformations, a two-way ANOVA on ranks was applied to detect differences in fish densities by size class among and within embayments (pooled transect site and monthly data). If significant differences were detected using the ANOVA test, then Dunn's multiple comparison procedure was used to detect differences in fish densities by size class among and within embayments.

A two-way ANOVA on ranks was used to detect differences in fish densities between fish length (size class) and habitats (pooled embayment, transect site, and monthly data). If significant differences were detected, then Dunn's multiple comparison procedure was used to identify differences in fish densities between specific fish lengths (size classes) and habitats.

Similarities of fish species present between habitats were measured using the Percent Similarity formula (PS) based on fish abundance (Gauch 1982). Transect site, month and embayment data were pooled for this comparison.

The four most common species observed in all northeast embayments were selected for further examination, For each of these species; monthly changes in fish densities for each embayment were compared. Monthly changes in densities of recruits ( $<5 \mathrm{~cm} \mathrm{TL}$ ) of each of these species were also compared.

For each key fish species, the distribution of fish density estimates for each embayment (pooled transect site and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If fish density estimates for each embayment failed the normality test even after $\log (x+1)$ and square root transformations, they were then compared using the nonparametric test Kruskall-Wallis One-Way ANOVA on ranks. Dunn's multiple comparison procedure was used to detect differences in mean densities of each key species between embayments.

Fish Traps - The distribution of fish and species trap catch rates estimates (pooled embayment, trap site and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If the distribution of fish and species trap catch rates for each embayment failed the normality test even after $\log$ ( x +1 ) and square root transformations, they were then compared using the non parametric test Kruskall-Wallis One-Way ANOVA on ranks. Dunn's multiple comparison procedure was used to detect differences in mean fish and species trap catch rates between embayments.

The four most common species caught in fish traps in all northeast embayments were selected for further examination. For each of these species, monthly changes in fish densities for each embayment were compared.

The distribution of trap catch rate estimates for each key species (pooled embayment, trap site and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If the distribution of trap catch rates for each key species (pooled trap site and monthly data) failed the normality test even after $\log (x+1)$ and square root transformations, they were then compared using the non parametric test Kruskall-Wallis One-Way ANOVA on ranks. Dunn's multiple comparison procedure was used to detect differences in mean trap catch rates between embayments for each key species.

Seine Nets - The distribution of fish and species seine haul catch rate estimates (pooled embayment, seine-haul and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). ). If fish and species seinehaul catch rates for each embayment failed the normality test even after $\log (x+1)$ and square root transformations, they were then compared using the non parametric test Kruskall-Wallis One-Way ANOVA on ranks. Dunn's multiple comparison procedure was used to detect differences in mean fish and species seine-haul catch rates between embayments.

Three key species caught in seine-hauls were identified. Changes in monthly catch rates of the most abundant species caught with seine net were compared by embayment.

For the three most abundant species, the distribution of seine-haul catch rate estimates (pooled embayment, seine-haul and monthly data) were checked for normality using the Kolmogorov-Smirnov normality test (SPSS Science 1997). If the distribution of seinehaul catch rates for each key species for each embayment failed the normality test even after $\log (x+1)$ and square root transformations, they were then compared using the non parametric test Kraskall-Wallis One-Way ANOVA on ranks. Dunn's multiple comparison procedure was used to detect differences in mean seine-haul catch rates between embayments for each key species.

## RESULTS

## Northeast Coast

## Visual Strip Transect Census

For Cottongarden Bay (pooled transect site and monthly data), a total of 4,471 fishes were observed representing 54 species and 21 families (see Table 1). The most abundant fish was the slippery dick, Halichoeres bivittatus (38\%) followed by the bucktooth parrotfish, Sparisoma radians ( $29 \%$ ). For Teague Bay (pooled transect site and monthly data), a total of 7,829 fishes were observed representing 66 species and 25 families (see Table 1). The most abundant genus was grunts, Haemulon unknown (38\%) followed by the slippery dick, H. bivittatus (13\%). Yellowcliff Bay (pooled transect site and monthly data), a total of 7,388 fishes were observed representing 74 species and 27 families (see Table 1). Small ( $<3 \mathrm{~cm} \mathrm{TL}$ ) newly settled grunts (Haemulidae) accounted for $33 \%$ of fish observed, Differentiation of these very small grunts to the species level was not possible. The slippery dick, H. bivittatus (27\%) was the second most abundant fish.

In Cottongarden Bay, seagrass was the dominant habitat sampled (Figure 3) covering almost $94 \%$ of the substrate found on transects. It was also the dominant habitat in Teague Bay transects, accounting for $84 \%$ of the habitat found. Yellowcliff Bay substrate was also dominated by seagrass ( $60 \%$ ). In Cottongarden Bay, Teague Bay, and Yellowcliff Bay patch reef accounted for $4 \%, 2 \%, 6 \%$ of the substrate surveyed respectively (Figure 3).

Peaks in monthly mean fish density were different for each embayment (Figures 4a). Monthly trends in species densities were similar for the three northern embayments (see Figure $4 b$ ). Species diversity ( $\mathrm{H}^{\prime}$ ) and evenness ( $\mathrm{J}^{\prime}$ ) mean values were similar for the three northern St. Croix embayments (see Figures 5a and b).

There were no significant differences in fish density, number of species per area, diversity and evenness indices between embayments (see Appendix B, numbers 1 to 4, $\mathrm{p}>0.05$ ),

For Teague Bay and Yellowcliff Bay, patch reef habitats had the highest number of fish and fish species per area compared with other habitats in those bays (see Figures 6a and
b). In Cottongarden Bay, rubble habitats had the highest fish and species densities compared with other habitats in this bay (Figures 6a and b).

Fish density and number of species per area estimates (pooled embayment, transect site and month data) for each habitat were significantly different (see Appendix B, numbers 5 and $6, \mathrm{p}<0.05$ ). Patch reef and rubble habitats had significantly higher fish densities and number of species per area than algal plains and sand habitats.

There were significant differences in $\mathrm{H}^{\prime}$ between habitats (Figure 7a, see Appendix B, number $7, \mathrm{p}<0.05$ ), but $\mathrm{J}^{\prime}$ showed no significant differences between habitats (Figure 7 b , see Appendix B, number 8, $\mathrm{p}>0.05$ ). H ' was significantly higher in patch reefs than other habitats (Figure 7a, Appendix B, number 7, $\mathrm{p}<0.05$ ).

There were significant differences in fish density between fish lengths (size classes) and habitat (Tables 2 and 3, and Appendix C, number 1, $\mathrm{p}<0.05$ ). Patch reefs and rubble habitats had significantly more fish $<5 \mathrm{~cm}$ TL than sand and seagrass habitats (Appendix C, number 2, $\mathrm{p}<0.05$ ). Density of larger fish ( $5-10 \mathrm{~cm}$ TL and $>10 \mathrm{~cm} \mathrm{TL}$ ) was significantly higher in patch reefs than in other habitats (Appendix C , numbers 3 and $4, \mathrm{p}<0.05$ ).

Among all northeast St, Croix embayments, divers counted more $<5 \mathrm{~cm}$ TL fish in Yellowcliff Bay. More larger fish ( $>10 \mathrm{~cm}$ TL) were counted in Teague Bay (see Table 3).

There were no significant differences in fish densities between fish lengths (size classes) and embayments using pooled transect site and monthly data (see Table 3, Appendix C, number $10, \mathrm{p}>0.05$ ).

Results of the percentage similarity of species composition and fish densities between habitats (see Table 4) indicated that algal plains and seagrass beds shared the greatest similarity ( $69 \%$ ) followed by seagrass and patch reefs ( $53 \%$ ). The sand habitat was very dissimilar (Table 4) and apparently were the preferred habitat of certain species such as the rosy razorfish, Hemipteronotus martinicensis, and the bridled goby, Coryphopterus glaucofraenum.

Based on the total number of economically important fish species observed (see Table 1), 17 economically important species from the St. Croix fishery were selected for subsequent analysis. A comparison of length frequencies for each of these seventeen economically important species by habitat is made in Table 5. Sparisoma viride, Scarus taeniopterus and Scarus iserti (Table 5) had higher densities of recruits ( $<5 \mathrm{~cm} \mathrm{TL}$ ) in patch reefs $\left(0.92,5.81\right.$, and 1.79 recruit $/ 100 \mathrm{~m}^{2}$, than juveniles and subadults ( $>10 \mathrm{~cm}$ TL) $\left(0.28,0.07\right.$, and 0 subadult $/ 100 \mathrm{~m}^{2}$, respectively). Ocvurus chrysurus recruit ( $<5 \mathrm{~cm}$ TL) densities (see Table 5) were highest on seagrass ( 0.41 recruit $/ 100 \mathrm{~m}^{2}$ ) while larger individuals ( $>10 \mathrm{~cm} \mathrm{IL}$ ) were most abundant on patch reefs ( 0.74 subadult/100 $\mathrm{m}^{2}$ ).

Densities of newly settled pants Haemulon unknown ( $<3 \mathrm{~cm} \mathrm{TL}$ ) were highest on rubble compared with other habitats (see Table 5). Fish counts for these small unidentified grunts were highest on seagrass. Seagrass represented the largest proportion of habitat surveyed here. Even though fish counts were higher in seagrass than rubble, density (per unit area) was lower.

The four most common species observed included: Haemulon unknown, H. bivittatus, $S$. radians and O. chrysurus (see Table 1). Monthly Haemulon unknown densities (for all fish size groups combined) appeared to peak in November 1998 and in June 1999 in Yellowcliff Bay (see Figure 8a). For Haemulon unknown recruits ( $<5 \mathrm{~cm}$ TL), density peaks were similar (compare Figures 8 a and $\mathfrak{b}$ ).

Monthly H. bivittatus densities (for all fish size groups combined) and recruit ( $<5$ cm TL) densities are presented in Figures 9 a and b. Density of $H$. bivittatus recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) at Yellowcliff Bay had a distinctive peak in April 1999.

Monthly S. radians densities (for all fish size groups combined) and recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) densities had a minor peak in October 1998 in all embayments (see Figures 10a and b). For Cottongarden Bay, both overall S. radians density and recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) density had a major peak in May 1999 (see Figures 10a and b).

Monthly $O$. chrysurus densities (for all fish size groups combined) and recruit ( $<5 \mathrm{~cm}$ TL) densities had small peaks in October 1998 and August 1999 for Teague Bay and Yellowcliff Bay, respectively (see Figures 1 1a and b). In Cottongarden Bay, overall O. chrysurus density peaked in April 1999, but O. chrysurus recruit ( $<5 \mathrm{~cm}$ TL) densities showed no similar patterns (compare Figures 11 a and b )

Densities of each of the four major species were compared by embayment. There was no significant difference in overall $O$. chrysurus densities between embayments (Appendix $B$, number $9, \mathrm{p}>0.05$ ). However, Cottongarden Bay had a significantly higher density of S. radians than the other embayments (Appendix B, number 10, $\mathrm{p}<0.05$ ); (2) the density of grunt, Haemulon unknown, recruits ( $<5 \mathrm{~cm}$ TL) in Yellowcliff Bay and Teague Bay were significantly higher than in Cottongarden Bay (Appendix B, number 11, $\mathrm{p}<0.05$ ); and (3) Yellowcliff Bay had higher overall density of the slippery dick, H. bivittatus than in Cottongarden Bay (Appendix B, number 12, $\mathrm{p}<0.05$ ).

Fish Trap - During this study period, 696 fish were caught by traps in Cottongarden Bay, These fishes represented 16 species and 12 families (Table 6). The spotted goatfish, Pseudupeneus maculatus, accounted for $38.6 \%$ of the total catch followed by the bucktooth parrotfish, S. radians (38.3\%). In Teague Bay, 760 fishes representing 26 species and 14 families were trapped (Table 6). In Teague Bay, the French grunt, $H$. flavolineatum ( $31 \%$ ) was the most commonly caught fish followed by the spotted goatfish, P. maculatus (19\%). For Yellowcliff Bay, a total of 438 fish representing 19 species and 11 families were trapped. The spotted goatfish, P.maculatus, was the most
common fish trapped (47\%) followed by the yellowtail snapper, O. chrysurus $(9 \%)$.
The monthly mean catch per trap-day (for all embayments) ranged from 1.8 to 21.8 fish per trap-day (Figure 12a). Teague Bay catch per trap-day peaked in October and November 1998, while that in Cottongarden Bay peaked in January 1999 (see Figure 12a).

Monthly number of species caught per trap-day (for all embayments) is presented in Figure 12b. In Teague Bay, the number of species caught per trap-day peaked in October 1998.

There were no significant differences in the number of fish and number of species caught per trap-day between embayments (Appendix B, numbers 13 and 14, $\mathrm{p}>0.05$ ).
H. flavolineatum, S. radians, P. maculatus and O. chrysurus were the most common species caught by trap in all embayments (Table 6). H. flavolineatum catches peaked in November of 1998 in Teague Bay (see Figure 13a). For this species, no peak was evident for the other two embayments. S. radians catch per trap-day was highest in Cottongarden Bay in January 1999 (Figure 13b). Monthly catch per trap-day for Yellowcliff and Teague Bays did not showed any peaks. Monthly catch per trap-day of $P$. maculatus was low from October 1998 to April 1999, but increased in all three embayments from June to September 1999 (Figure 13c). Yellowtail snapper, O. chrysurus, monthly catch per trapday peaked in October 1998 in Teague Bay and in December of 1999 in Yellowcliff Bay (Figure 13d).

For fish caught by trap, length (size class) frequency distributions of $H$. flavolineatum, 0 . chrysurus, P. maculatus and S. radians show that trap sampling caught small fish (Figures 14 to 17). For reference, mean fish sizes at sexual maturity (length at which $50 \%$ of the population become mature for the first time) are noted in each of these figures. Results here suggest that in general, fish trapped were smaller than their mean size at sexual maturity (Munro 1983; Claro 1983; and Billings and Munro 1974).

There were no significant differences in catch per trap-haul between embayments for H. flavolineatum, 0. chrysurus, and P. maculatus (Appendix B, numbers 15 to 17, $\mathrm{p}>0.05$ ). However, Cottongarden Bay had a significantly higher mean number of S . radians caught per trap-day than Teague Bay (Appendix B, number 18, $\mathrm{p}<0.05$ ).

Monthly changes in length (size class) frequency distribution of key fish species were not compared here because of the small number of fish caught each month.

Beach Seine - During this study period, 61 fish were caught by seine net in Cottongarden Bay. These fishes represented 9 species and 7 families (Table 7), The permit, Trachinotus falcatus, accounted for $26 \%$ of beach seine catches, followed by the slender mojarra, Eucinostomus jonesi (19\%). In Teague Bay, 263 fishes representing 29 species and 19 families were caught. The horse-eye jack, Caranx latus accounted for $28 \%$ of the seine
catches followed by the slender mojarra, E. jonesi (19\%). At Yellowcliff Bay, 10 fishes representing 4 species and 4 families were caught. The slender mojarra, E. jonesi, and the hound fish, Tylosurus crocodilus, dominated the species composition ( $50 \%$, and $30 \%$, respectively) of total fish caught by seine net (Table 7).

Mean monthly variations in fish caught per seine-haul for the three bays ranged from 0 to 29 fish/seine-haul (Figure 18a). In Teague Bay, CPUE was highest in May 1999 (29 fishes per seine-haul).

Mean monthly variations in species caught per seine-haul varied from 0 to 6 species per seine-haul (see Figure 18b). For mean monthly number of species per haul the largest peak was in March 1999 in Teague Bay (see Figure 18b).

There were significant differences in the number of fish and number of species caught per seinehaul (based on pooled seine-haul, monthly, and embayment data) between embayments (Appendix B, numbers 19 and 20, $\mathrm{p}<0.05$ ). Teague Bay had significantly more fishes and species per seine-haul than any of the other sites (Appendix B, numbers 19 and 20, p<0.05, Figure 18b).
C. latus, T. falcatus and E. jonesi were the three most abundant species caught by seine nets in all embayments (Table 7). In Teague Bay, C. latus had a major catch per seinehaul peak in April 1999 (Figure 19a). In Teague and Cottongarden Bays, E. jonesi catch per seine-haul peaked in May 1999 (see Figure 19b). For T. falcatus, catch per seinehaul was highest from March to August 1999.

Length-frequency distributions of C. latus, T. falcatus and E. jonesi from seine hauls for all embayments are presented in Figures 20a to g. For reference, mean fish sizes at sexual maturity are noted in each of these figures. Results here suggest that fish caught by seinehaul were smaller than their mean size at sexual maturity (Thompson and Munro 1974; Garcia-Cagide et al 1994)

There was no significant difference in the number fish caught per seine-haul for each key species (C. latus, T. falcatus and E. jonesi), between embayments based on pooled seinehaul, monthly, and embayment data (Appendix B, numbers 21 to 23, $\mathrm{p}>0.05$ ).

## Southeast Coast

## Visual Strip Transect Census

During this survey period, a total of 1,628 fishes representing 39 species and 16 families were observed in Turner Hole (Table 8). The most abundant fish observed was the bluehead wrasse, Thalassoma bifasciatum (15\%) followed by newly settled grunts, Haemulon unknown (13\%). In Robin Bay, 309 fishes from 25 species and 14 families were observed during this survey period. In Robin Bay, the slippery dick, H. bivittatus, accounted for $30 \%$ of fish observed, followed by the bucktooth, S. radians ( $25 \%$ of fish observed), At Great Pond Bay, 1,285 fish representing 36 species and 12 families were
observed during this survey. Newly settled ( $<3 \mathrm{~cm} \mathrm{TL}$ ) grunts (Haemulidae) accounted for $69 \%$ of the total fish observed abundance followed by the slippery dick, H. bivittatus (8\%).

In Turner Hole, seagrass and sand accounted for $66 \%$ and $21 \%$ respectively of the habitat observed in transects (Figure 21). In Robin Bay, seagrass accounted for almost $92 \%$ of the substrate. Great Pond Bay was dominated by sand (75\%). In Turner Hole, Robin Bay, and Great Pond Bay patch reef accounted for $5 \%, 0.5 \%$ and $0.7 \%$ of the substrate surveyed respectively (Figure 21).

There were significant differences in fish density, number of species per area (pooled transect site, monthly, and embayment data), diversity index and evenness index estimates between embayments (see Appendix D, numbers 1 to $4, \mathrm{p}<0.05$ ). Turner Hole had a significantly higher (1) fish density, (2) number of species per area, and (3) diversity $\left(\mathrm{H}^{\prime}\right)$ than other embayments. Evenness ( $\mathrm{J}^{\prime}$ ) values for Great Pond were higher than those of other embayments (Appendix D, numbers 1 to $4, \mathrm{p}<0.05$ ),

For habitats and embayments (using pooled transect site and monthly data), fish density (Figure 22a) and number of species per area (Figure 22b) were highest in patch reef habitats. There were significant differences in fish densities and number of species per area between habitats using pooled transect site, monthly, and embayment data (see Appendix D, numbers 5 and $6, \mathrm{p}<0.05$ ). Patch reef and algal plain habitats had significantly higher fish densities than seagrass beds and sand habitats (see Figure 22a, Appendix D, number 5, $\mathrm{p}<0.05$ ). Patch reef habitats had significantly higher number of species per area than seagrass and sand (see Figure 22b, Appendix D, number 6, p<0.05). Rubble and algal plain had significantly higher number of species per area than sand (see Appendix D, number 6, $\mathrm{p}<0.05$ ).

There were significant differences in diversity index $\left(\mathrm{H}^{\prime}\right)$ and evenness index ( $\mathrm{J}^{\prime}$ ) estimates (pooled transect site, monthly, and embayment data) between habitats (see Appendix D, numbers 7 and $8, \mathrm{p}<0.05$ ). $\mathrm{H}^{\prime}$ of patch reefs and rubble habitats had significantly higher values than that of sand and algal plain (see Figure 23a, Appendix D, number 7, $\mathrm{p}<0.05$ ). Seagrass beds and algal plain had significantly higher evenness J' indices compared with sand habitats (Figure 23b, Appendix D, number 8, p<0.05).

Patch reefs and algal plain habitats had significantly higher recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) densities than sand and seagrass habitats (Appendix E, number 1 and 2, p $<0.05$ ). Fish densities of $5-10 \mathrm{~cm}$ TL fish were significantly higher on patch reefs than seagrass, sand and algal plain habitats (Appendix E, number 3, p<0.05). Fish densities of fish $>10 \mathrm{~cm}$ TL were significantly higher on patch reefs than in all other habitats except algal plain habitats (Appendix E, number 4, $\mathrm{p}<0.05$ ).

Among all southeast St. Croix embayments, Great Pond Bay Bay had more $<5 \mathrm{~cm}$ TL fish counted, while Turner Hole Bay had more larger fish ( $>10 \mathrm{~cm} \mathrm{TL}$ ) counted (Table 10).

Turner Hole and Great pond had significantly higher densities of recruits ( $<5 \mathrm{~cm} \mathrm{TL}$ ) than Robin Bay (Appendix E, number 10 and 11, $\mathrm{p}<0.05$ ). However Turner Hole had significantly higher densities of fish $5-10 \mathrm{~cm}$ TL and $>10$ TL than in other embayments (Appendix E, numbers 12 and $13, \mathrm{p}<0.05$ ).

Results of the percentage similarity of species composition and fish densities between habitats (see Table 11) indicated that patch reefs and rubble had the greatest similarity of species ( $88 \%$ ) followed by algal plain and sand (72\%).

Based on the total number of species observed (see Table 8), the top seventeen economically important species were selected. A comparison of length frequencies for each of these seventeen economically important species by habitat is made in Table 12. S. viride, and S. taeniopterus had higher recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) densities in patch reefs ( 3.35 , and 4.47 recruit $/ 100 \mathrm{~m}^{2}$, respectively) than conspecific juveniles and subadults $>10$ cm TL (see Table 12). O. chrysurus recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) densities were highest on seagrass ( 0.39 recruit $/ 100 \mathrm{~m}^{2}$ ) while larger individuals ( $>10 \mathrm{~cm} \mathrm{TL}$ ) were primarily found on patch reefs ( 0.84 subadult $/ 100 \mathrm{~m}^{2}$ ). In addition, algal beds had higher newly settled grunts densities ( 61.08 newly settled grunts $/ 100 \mathrm{~m}^{2}$ ), than patch reefs ( 25.98 newly settled grunts $/ 100 \mathrm{~m}^{2}$ ) and sand habitats ( 3.08 newly settled grunts $/ 100 \mathrm{~m}^{2}$ ). Densities for larger size groups ( $>10 \mathrm{~cm}$ TL) for H. flavolineatum and H. plumieri were highest on patch reefs ( 2.79 and 3.79 subadults $/ 100 \mathrm{~m}^{2}$, respectively) than in other habitat.

The four most common species observed included: Haemulon unknown, H. bivittatus, S. radians and T. bifasciatum (see Table 8). There were no significant differences in fish densities between embayments for Haemulon unknown, S. radians, and T. bifasciatum (Appendix D, numbers 9 to 11, $\mathrm{p}>0.05$ ). However, for H. bivittatus, there was a significant difference between fish densities between embayments. Turner Hole Bay had a significantly higher density of $H$. bivittatus compared with Robin Bay (Appendix D, number $12, \mathrm{p}<0.05$ ),

Fish Trap - In Turner Hole Bay, 67 fishes were caught by traps. These fishes represented 13 species and 10 families (Table 13). The most common fish caught was the bucktooth parrotfish, S. radians, accounting for $38 \%$ of the total fish caught. At Robin Bay, 38 fishes representing 7 species and 6 families were caught by traps. The bucktooth parrotfish, S. radians, accounted for $63 \%$ of the total catch at Robin Bay. At Great Pond Bay, 12 fishes representing 5 species and 5 families were caught by trap. The doctorfish, Acanthurus chirurgus, accounted for $33 \%$ of the catch at Great Pond.

There was no significant difference in the number of fish caught per trap-haul and number of species caught per trap-haul between embayments (Appendix D, numbers 13 and $14, \mathrm{p}>0.05$ ).
A. chirurgus, $S$. radians, $P$. maculatus and $O$. chrysurus were the most common species
caught in traps at all embayments (Table 13). Due to the small sample size and duration (only three months), differences in fish catch rates between embayments and months could not be examined.

Length-frequency distributions of key fish species caught in traps are presented in Figures 24a to d. Mean Size at sexual maturity for each of these key species (if available) was noted in Figures 24a to d. Peaks in length frequency for all key species caught were below the mean length for sexual maturity of those species (Claro 1983; and GarciaCagidae et al 1994).

Monthly changes in length frequency distribution of key fish species were not compared here because of the small number of fish caught each month.

Beach Seine - In Turner Hole, 9 fishes were caught by seine net representing 3 species and 2 families (Table 14). The permit, T. falcatus, accounted for $77 \%$ of Turner Hole seine catches. In Robin Bay, 10 fishes were caught representing 5 species and 5 families. The permit, T. falcatus, accounted for $50 \%$ of the catches in Robin Bay. At Great Pond Bay, 20 fishes representing 7 species and 6 families were caught. The herring, Jenkinsia lamprotaenia, and the school master, Lutjanus apodus, accounted for $35 \%$, and $30 \%$ of the catch, respectively (Table 14).

There were no significant differences in the number of fish and number of species caught per seine-haul between embayments (Appendix D, numbers 15 and 16, $\mathrm{p}>0.05$ ).

The two most common species caught in seine-hauls at all embayments were T. falcatus and L. apodus (Table 14). Due to limited data collected (only three months), variations in seine-haul catches between months and embayments could not be examined.

Length-frequency distributions of T. falcatus and L. apodus (using pooled seine-haul sets, monthly, and embayment data) are presented in Figures 25a and b. Mean size at sexual maturity for each of these key species was noted in Figures 25a and b. Peaks in length frequency for these two key species were below the mean length of sexual maturity of those species (Thompson and Munro 1983; Garcia-Cagide et al 1994).

## DISCUSSION

In this study, many fish were found to utilize nearshore habitats as nursery areas, including Haemulon unknown, H. aurolineatum, H. flavolineatum and S. taeniopterus (see Table 5 and 12). S. radians, H. bivittatus and Haemulon unknown, accounted for
$>65 \%$ of all fishes observed by divers in the visual strip transect census in northeast St. Croix embayments. In southeast St. Croix embayments, H. aurolineatum, $H$. bivittatus, S. radians and Haemulon unknown accounted for more than $55 \%$ of all fishes observed by divers in the visual strip transect census.

Results of this study on St. Croix backreef lagoonal areas can be related to results from similar studies in other lagoonal areas (Ogden and Zieman 1977; Gladfelter and Gladfelter 1978; Brothers and McFarland 1981; Thayer et al 1987; Baelde 1990; and Dennis 1992). Comparisons of the ichthyofauna of the St. Croix bank barrier reef lagoons with other lagoon systems studied in the Caribbean indicate structural similarities, with some marked differences. Stoner (1986) found that soleids and gerreids dominated a mangrove lagoon system in Joyuda, Puerto Rico. However, haemulids, scarids, and labrids (the most abundant fishes found in this study) were a minor component. There were no mangrove areas surveyed in this study. Yanes-Arancibia et al (1980), Thayer et al (1987), and Sogard et al (1989) found that engraulids, clupeids and gerreids dominated the species composition in Terminos Lagoon in western Yucatan and in seagrass beds adjacent to Florida Bay mangrove system. Baelde (1990) found that gerreids, engraulids and sparids dominated the T. testudinum beds in Guadulope.

Studies done on lagoons close to bank barrier reef lagoons showed distinct reef fish communities dominated by scarids, haemulids, and labrids. Sedberry and Cartier (1993) found that Haemulidae such as Haemulon sciurus and H. flavolineatum dominated the species composition in a survey of fish communities in barrier reef lagoons in Belize. In a Tortola study (Mackey 1999), juvenile fish assemblages on barrier reef lagoons were dominated by H. aurotineatum and H. flavolineatum. In Curazao, Nagelkerken et al (2000) found H. flavolineatum and S. iserti were the dominant species of reef fish in barrier reef lagoons.

The differences in dominant species between these studies may be related to physical conditions and proximity to the main reef. The lagoons sampled by Stoner (1986), Yanez-Aranciba et al (1980), Thayer et al (1987) Sogard et al (1989) and Baelde (1990) had more restricted access to open water, had more estuarine conditions (lower salinity), and did not have extensive nearby barrier reef formations. This was the case in Belize (Sedberry and Cartier 1993), Tortola (Mackey 1999), Curazao (Nagelkerken et al 2000) and St. Croix (Ogden and Zieman 1977; and Gladfelter and Gladfelter 1978, Robblee and Zieman 1984; and Shulman 1984).

In the Caribbean, the interaction of seagrass beds and mangrove ecosystems with coral reef fish species constitutes an additional factor influencing the structure of shallow water fish communities (Robblee and Zieman 1984; and Gladfelter and Gladfelter 1978). The relationship between seagrass beds and coral reefs is well documented, in particular the foraging behavior of coral reef fishes over seagrass beds (Ogden and Zieman 1977; Robblee and Zieman 1984; and Shulman 1984).

The more abundant fish species found in this study were primarily herbivores and
planktivores (see Table 1 and 9). Some of these dominant species are recreationally and commercially important to the local nearshore fishery. Most fish (86\%) inhabiting the backreef lagoon areas in the northeast coast embayments (Cottongarden Bay, Teague Bay and Yellowcliff Bay) were small ( $<10 \mathrm{~cm}$ TL). Very few (14\%) embayment fish observed were over 10 cm TL. This is also true of the southeast embayments (Turner Hold Bay, Robin Bay, and Great Pond Bay) where there were few fish (7\%) over 10 cm TL. Results here indicated that backreef lagoon areas play an important role as a nursery grounds for economically important species in the early stages of their life cycle.

Previous studies on Teague Bay bank barrier reef lagoon have shown the importance of these areas for juvenile reef fishes (Shulman 1984; McFarland et al 1985; and Shulman and Ogden 1987). According to Shulman and Ogden (1987), post settlement mortality of newly settled $H$ flavolineatum recruits was high (90\%). Despite this, Shulman and Ogden (1987) reported that the mean annual recruitment at Teague Bay bank barrier reef lagoon was among the highest rates for this species yet reported for reef fishes $\left(44 / \mathrm{m}^{2}\right)$. In that study. H. flavolineatum recruit abundance increased with increasing distance from the reef They attributed this to the large number of juvenile grunts utilizing shallow lagoonal habitats. According to Shulman and Ogden (1987), the number of predators decreased with increased distance from the reef as the number of predators in shallow lagoon waters was lower than on reefs. Additional protection of recruits was also provided by water turbidity and seagrass shoots. Spatial separation of juveniles from adults that reside on the reef also reduced intraspecific competition and could increase survival rates during recruitment (Shulman and Ogden 1987).

At the sites studied here, fish species formed distinct communities within seagrass, sand, patch reefs, algal plains and rubble habitats. The seagrass and algal plain habitats at all embayments were dominated by small resident fish such as Halichoeres spp., S. radians and juveniles of non-resident, economically important species such as Haemulon unknown. and $O$. chrysurus. Rubble and patch reef habitats harbored a higher number of species per unit area and were mostly dominated by small juvenile damselfish, parrotfish, grunts, and doctorfish.

The differences in fish size distributions of some dominant economically important species between habitats suggested different habitat preferences by those fish species. Many local economically important coral reef fish (S. taeniopterus, S. iserti, Acanthurus chirurgus, and $A$. bahianus) recruits were found on patch reefs, rubble and seagrass habitats. Species such as newly settled grunts Haemulon unknown, $O$. chrysurus and Lutjanus synagris recruits were mostly found in seagrass beds (Tables 5 and 12). Densities of larger ( $>10 \mathrm{~cm} \mathrm{TL}$ ) fish (S. taeniopterus, S. iserti and $O$. chrysurus) were low on patch reefs and seagrass beds (Tables 5 and 12). These results may reflect temporary and successive utilization of seagrass beds and patch reefs by juveniles of various species that move elsewhere as they grow larger. This was the case for $O$. chrysurus, S. taeniopterus and S. iserti.

The present study suggests that nursery habitats are not limited to mangrove habitats, but
include other nearshore habitats such as patch reefs and rubble areas. In this study, patch reefs and rubble habitats had the highest densities of fish recruits. However, total counts of fish recruits were higher on seagrass beds and algal plains. This was because of the dominance of these habitats in embayments surveyed (Table 2 and Table 9). Recent Caribbean studies of nearshore fish assemblages suggest that patch reefs and rubble areas appear to be important shelter sites for juvenile fishes in mangrove and seagrass dominated lagoons (Risk 1997, Nagelkerkern et al 2000, Adams and Ebersole, in press).

Seasonal changes in the density of fishes and the number of species per area in northeast St. Croix embayments were observed in most cases (Figures 4 a and b, and 8a to 11b). Previous studies done at Teague Bay (Rogers and Salesky 1981; and Carpenter 1986) found that during the waimer months (from July to November) there was a significant increase in algal cover and biomass that probably stimulated increases in the invertebrate population. This increase in food availability may favor herbivores (such as $S$. radians, $S$. iserti, S. taeniopterus), omnivores (such as H. bivittatus, Haemuion unknown) and small carnivorous fishes such as O. chrysurus.

Seasonal variations may also be related to the reported correlation between increased larvae/juvenile fish settlement and recruitment and summer increases in the number of species and fish present (Williams and Sale 1981; and Doherty and Williams 1988). In this study, summer recruitment pulses were evident for some species such as Haemulon unknown although the magnitude of these pulses varied between embayments (Figure 8 b ).

Recent studies in St. Croix on larval fish supply and microhabitat characteristics have had contrasting results (Caselle and Warner 1996; Tolimieri 1995, 1998a and b; Tolimieri et al. 1998; Risk 1997; and Nemeth 1997). Caselle and Warner (1996) found that microhabitat characteristics were correlated with the abundance of T. bifasciatum, but did not explain recruitment of $T$. bifasciatum at all sites around St , Croix. On the north windward shore, recruitment rates increased from east to west while on the south leeward side, recruitment rates increased from west to east. They concluded that along the two shores of St. Croix, physical transport processes were more important determinants of the spatial patterns of recruitment than habitat selection. In contrast, Tolimieri (1995) could not show any microhabitat selection during recruitment of Stegastes planifrons.

Microhabitat use may be an important factor in establishing consistent patterns of recruitment. Tolimieri (1998a and b) found that S . viride and $S$. iserti recruitment rates at Teague Bay were correlated with microhabitat characteristics at a smaller spatial scale. Tolimieri et al (1998) studied recruitment patterns of different reef fish species throughout the north coast of St. Croix.

They concluded that some of the recruitment patterns of T. bifasciatum could be explained by larval transportation. However, the recruitment rates for S. viride, S. planifrons, Scarus spp. and Stegastes spp. did not conform to the predicted spatial recruitment patterns stated by Caselle and Warner (1996), This suggested that other factors such as habitat selection may be influencing recruitment patterns for some species.

Recruitment studies for the ocean surgeonfish, A. bahianus, and the bicolor damselfish, Stegastes partitus, at Teague Bay (Risk 1997 and Nemeth 1997) suggest that there are several post-settlement factors at work. These factors may include the ability of postsettlement stage fishes to select habitat, the level of competition for food, and spatial distribution of resources. Larval supply alone appears to be insufficient to explain the abundance and distribution of reef fish populations in St. Croix (Risk 1997 and Nemeth 1997).

Lower overall fish recruitment observed in Cottongarden Bay may be due to lower larval supply but also to a shortage of suitable habitat. Haemulon unknown recruits were rare in Cottongarden Bay. This may be due to the homogeneity of the habitat (almost $95 \%$ seagrass, see Figure 3). In this embayment, suitable habitats for fish recruits such as patch reefs, algal plain, and rubble
areas were scarce. Yellowcliff Bay had more complex habitat composition and harbored twice the number of fish recruits observed as Cottongarden Bay (Table 3),

The three methods used to sample nearshore habitats yield satisfactorily although somewhat different results due to biases inherent in each method (Boulon 1992). Fish traps tended to under represent certain species (such as H. bivittatus) that apparently avoided traps or were small enough to pass through the trap mesh and not to be caught (Mackey 1999). However, traps allowed accurate total length measurements of the most abundant species. Baitfish such as the false pilchard, small permits and slender mojarras were more abundant in seine net catches than in transects and fish traps. However, seines were limited to areas adjacent to the shore without rocky and hard bottom substrates. The visual census method can provide a list of species in an area. It is possible to collect some length frequency information based on visual estimates of fish size. However, visual estimates are not as accurate as actually measuring each fish. Nevertheless, by utilizing these three methods at the same time, a more complete view of the finfish community can be achieved than from any one method alone.

It is apparent from this study that the backreef lagoon areas of Cottongarden Bay, Teague Bay Yellowcliff Bay, Turner Hole Bay, Robin Bay, and Great Pond Bay are important nurseries for many economically important fishes. While variability among sites and months were documented, general trends were similar. Results here document the importance of these nearshore habitats for juvenile fishes. Most fish sampled by trap and beach seine nets were smaller in size then their mean size at sexual maturity (for trap catches see Figures 14 to 17 and 24a to d, for beach seine catches see Figures 20a to land $25 a$ and b), Since the majority of juvenile fishes in nearshore habitats are economically important species, these habitats must be conserved to ensure the continued viability of the fisheries in St. Croix.

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Table 1. Total Number of Fish and Species Observed, and Percentage (Number of a Fish Species Observed/Total Number of Fish Observed in an Embayment) for Each Northeast St. Croix Embavment ( 120 Transects Per Embavment) from October 1998 and September 1999 Based on Pooled Monthly and Transect Site Data.

| FAMILY | SPECIES | Cottongarden Bay |  | Teague Bay |  | Yellowcliff Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish | Percent | No. of Fish | Percent | No. of Fish | Percent |
| Acanthuridae | Acantharus bahianus | 61 | 1.36 | 16 | 0.20 | 34 | 0.46 |
|  | Acanthurus chirurgus | 120 | 2.68 | 64 | 0.82 | 281 | 3.80 |
|  | Acanthurus coeruleus | 41 | 0.92 | 3 | 0.04 | 26 | 0.35 |
| Apogonidae | Apogon unk. | * | * | * | $\cdots$ * | 3 | 0.04 |
| Aulostomidae | Aulostomus maculatus | 3 | 0.07 | 1 | 0.01 | * | * |
| Balistidae | Balistes vetula | * | $\cdots$ * | * |  | 1 | 0.01 |
| Blenidae | Labrisomus unk. | * | * | 1 | 0.01 | * | * |
| Bothidae | Bothus lunatus | * | * | 1 | 0.01 | * |  |
| Carangidae | Caranx bartholomaei | 1 | 0.02 | * | * | 2 | 0.03 |
|  | Caranx crysos | 1 | 0.02 | 8 | 0.10 | 7 | 0.09 |
|  | Caranx ruber | 7 | 0.16 | 7 | 0.09 | 11 | 0.15 |
|  | Selar crumenophthalmus | * | * | 400 | 5.11 | * | * |
| Chaetodontidae | Chaetodon capistratus : | 18 | 0.40 | 34 | 0.43 | 91 | 1.23 |
|  | Chaetodon ocellatus | * | - * | * | * | 1 | 0.01 |
|  | Chaetodon striatus | 1 | 0.02 | 2 | 0.03 | 2 | 0.03 |
| Dactylopteridae | Dactylopierus volitans | . * | * | 1 | 0.01 | * |  |
| Dasyatidae | Dasyatis americana | - * | * | 1 | 0.01 | * | * |
| Diodontidae | Diodon hystrix | * | * | * | * | 2 | 0.03 |
| Engraulidae | Anchoa byolepis | * | * | * | * | 200 | 2.71 |
| Gerreidae | Gerres cinereus | 1 | 0.02 | 3 | 0.04 | 1 | 0.01 |
|  | Eucinostomus jonesi |  |  | 14 | 0.18 | * | * |
| Gobiidae | Gobiosoma unk. | 1 | 0.02 | 1 | 0.01 | 5 | 0.07 |
|  | Coryphopterus glaucofraenum | * | * | 7 | 0.09 | 13 | 0.18 |
|  | Gobiosoma genie | * | * | 20 | 0.26 | * | * |
|  | Coryphopterus unk. | * | * | * | * | 7 | 0.09 |
|  | Gnatholepis thompsoni | * | * | * | * | 1 | 0.01 |
| Gramistidae | Rypticus saponaceus | * | * | 1 | 0.01 | * | * |
| Haemulidae | Haemulon unk. | 45 | 1.01 | 2983 | 38.10 | 2465 | 33.36 |
|  | Haemulon aurolineatum | 6 | 0.13 | 921 | 11.76 | 210 | 2.84 |
|  | Haemulon carbonarium | 13 | 0.29 | * | * | 3 | 0.04 |
|  | Haemulon chrysargyreum | 5 | 0.11 | * | * | * |  |
|  | Haemulon flavolineatum | 87 | 1.95 | 471 | 6.02 | 109 | 1.48 |
|  | Haemulon plumieri | 129 | 2.89 | 41 | 0.52 | 10 | 0.14 |
| Pomacandithae | Holacanthus ciliaris | * | * | * | * | 2 | 0.03 |
| Holocentridae | Holocentrus adscensionis | 15 | 0.34 | 8 | 0.10 | 37 | 0.50 |
|  | Myripristis jacobus | 12 | 0.27 | 5 | 0.06 | 4 | 0.05 |
| Labridae | Doratonotus megalepsis | 1 | 0.02 | 3 | 0.04 | * |  |
|  | Halichoeres unk. | 7 | 0.16 | 20 | 0.26 | 33 | 0.45 |
|  | Halichoeres bivittatus | 1687 | 37.73 | 1058 | 13.51 | 1967 | 26.62 |
|  | Halichoeres garnoti | 12 | 0.27 | 33 | 0.42 | 58 | 0.79 |
|  | Halichoeres maculipinna. | 7 | 0.16 | * | * | * |  |
|  | Halichoeres poeyi | 84 | 1.88 | 48 | 0.61 | 75 | 1.02 |
|  | Halichoeres radiatus | * | * | * | * | 1 | 0.01 |

Table 1 (Continued). Total Number of Fish and Species Observed, and Percentage (Number of a Fish Species Observed/Total Number of Fish Observed in an Embayment) for Each Northeast St. Croix Embayment (120 Transects Per Embayment) from October 1998 and September 1999 Based on Pooled Monthly and Transect Site Data.

| FAMLY | SPECIES | Cottongarden Bay |  | Teague Bay |  | Yellowcliff Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish | Percent | No. of Fish | Percent | No. of Fish | Percent |
| Labridae | Hemipteronotus martinicensis | * | * | 156 | 1.99 | 197 | 2.67 |
|  | Hemipteronotus splendens | * | \% * | * | * | 8 | 0.11 |
|  | Thalassoma bifasciatum | 218 | 4.88 | 11 | 0.14 | 44 | 0.60 |
| Lutjanidae | Lutjanus apodus | 5 | 0.11 | 1 | 0.01 | * | * |
|  | Lutjanus griseus | * | * | 2 | 0.03 | 3 | 0.04 |
|  | Lutjanus mahogoni | 3 | 0.07 | 7 | 0.09 | 14 | 0.19 |
|  | Lutjanus synagris | 1 | 0.02 | 14 | 0.18 | 4 | 0.05 |
|  | Ocyurus chrysurus | 115 | 2.57 | 221 | 2.82 | 162 | 2.19 |
| Monacanthidae | Monacanthus ciliatus | 14 | 0.31 | 7 | 0.09 | 9 | 0.12 |
| Malacanthidae | Malacanthus plumieri | * | * | * | * | 1 | 0.01 |
| Mullidae | Mulloidicthys martinicus | * | * | 203 | 2.59 | * | * |
|  | Pseudupeneus maculatus | 113 | 2.53 | 26 | 0.33 | 76 | 1.03 |
| Muraenidae | Gymnothorax moringa | 1 | 0.02 | * | * | * | * |
| Ophichthidae | Ophichthus spp. | * | * | 5 | 0.06 | * | * |
|  | Ophichthus ophis | * | * | * | * | 1 | 0.01 |
| Ostracidae | Lactophrys polygonia | * | * | 1 | 0.01 | * | * |
|  | Lactophrys triqueter | 1 | 0.02 | * | * | 1 | 0.01 |
| Paralichtydae | Paralichthys tropicus | * | * | * | * | 1 | 0.01 |
| Pomacanthidae | Pomacanthus para | * | * | * | * | 2 | 0.03 |
| Pomacentridae | Abudefduf saxatilis | * | * | * | * | 3 | 0.04 |
|  | Microspathodon chrysurus | 15 | 0.34 | 5 | 0.06 | 10 | 0.14 |
|  | Stegastes unk. | * | - * | * | * | 12 | 0.16 |
|  | Stegastes diencaeus | * | * | * | * | 1. | 0.01 . |
|  | Stegastes fuscus | * | * | 21 | 0.27 | 55 | 0.74 |
|  | Stegastes leucostictus | 2 | 0.04 | 191 | 2.44 | 153 | 2.07 |
|  | Stegastes partitus | $\because 19$ | 0.42 | 11 | 0.14 | 79 | 1.07 |
|  | Stegastes planifrons | 2 | 0.04 | 3 | 0.04 | 25 | 0.34 |
|  | Stegastes variabilis | 12 | 0.27 | 5 | 0.06 | 21 | 0.28 |
| Scaridae | Cryptotomus roseus | 58 | 1.30 | 53 | 0.68 | 82 | 1.11 |
|  | Scarus unk. | * | * | * | * | 17 | 0.23 |
|  | Scarus iserti | 41 | 0.92 | 42 | 0.54 | 46 | 0.62 |
|  | Scarus taeniopterus | 138 | 3.09 | 94 | 1.20 | 89 | 1.20 |
|  | Sparisoma aurofrenatum | 13 | 0.29 | -* | * | 27 | 0.37 |
|  | Sparisoma radians | 1289 | 28.83 | 496 | 6.34 | 443 | 6.00 |
|  | Sparisoma rubripime | 3 | 0.07 | * | * | * | * |
|  | Sparisoma viride | * | * | 6 | 0.08 | 49 | 0.66 |
|  | Sparisoma chrysopterum | * | * | * | * | 3 | 0.04 |
| Sciaenidae | Equetus acuminatus | 2 | 0.04 | * | * | 2 | 0.03 |
| Scombridae | Scomberomonus regalis | 1 | 0.02 | * | * | * | * |
| Serranidae | Alphestes afer | 12 | 0.27 | 1 | 0.01 | * | * |
|  | Epinephelus guttatus | 10 | 0.22 | 5 | 0.06 | 11 | 0.15 |

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Table 1 (Continued). Total Number of Fish and Species Observed, and Percentage (Number of a Fish Species Observed/Total Number of Fish Observed in an Embayment) for Each Northeast St. Croix Embayment (120 Transects Per Embayment) from October 1998 and September 1999 Based on Pooled Monthly and Transect Site Data.

| FAMILY | SPECIES | Cottongarden Bay |  | Teague Bay |  | Yellowcliff Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish | Percent | No. of Fish | Percent | No. of Fish | Percent |
| Serranidae | Epinephelus striatus | * | * | 1 | 0.01 | * | * |
|  | Hypoplectrus spp | * | * | 1 | 0.01 | 2 | 0.03 |
|  | Hypoplectrus nigricans | * | * | 1 | 0.01 | * | * |
|  | Hypoplectrus indigo |  | * | 1 | 0.01 | * | * |
|  | Hypoplectrus puella | * | * | * | * | 1 | 0.01 |
|  | Hypoplectrus unicolor | * | * | 1 | 0.01 | 1 | 0.01 |
|  | Serramus tabacarius | * | * | 1 | 0.01 | * | * |
|  | Serranus tigrimus | 6 | 0.13 | 5 | 0.06 | 20 | 0.27 |
| Sphyraenidae | Sphyraena barracuda | * | * | * | * | 2 | 0.03 |
| Synodontidae | Synodus spp. | 1 | 0.02 | 9 | 0.11 | 3 | 0.04 |
|  | Synodus foetens | * | * | 1 | 0.01 | 3 | 0.04 |
| Tetradontidae | Canthigaster rostrata | 10 | 0.22 | 27 | 0.34 | 30 | 0.41 |
|  | Sphoeroides spengleri | 1 | 0.02 | 18 | 0.23 | 13 | 0.18 |
|  | Total Number of Fish | 4471 |  | 7829 |  | 7388 |  |
|  | Total Number of Species | 54 |  | 66 |  | 74 |  |

Mean (Standard Error) Community Parameters of fish censused with visual transects during October 1998 to September 1999. All months were pooled.

|  | Cottongarden Bay | Teague Bay | Yellowcliff Bay |
| :--- | :---: | :---: | :---: |
| No. of fish $/ 100 \mathrm{~m}^{2}$ | $19.51(1.99)$ | $32.75(5.37)$ | $30.43(4.81)$ |
| No of species $/ 100 \mathrm{~m}^{2}$ | $2.93(0.21)$ | $2.95(0.18)$ | $3.48(0.23)$ |
| Diversity Index | $0.63(0.04)$ | $0.58(0.04)$ | $0.67(0.04)$ |
| Evenness | $0.70(0.02)$ | $0.66(0.02)$ | $0.65(0.02)$ |

*none observed

Table 2. Number of Fish Observed and Number of Fish Densities By Size Groups and By Habitat Based on Pooled Monthly, Transest. Site-and.Northeast St. Cenix_Fmhavment Data ( C ctoher 1998 to Sentemher 1999),

| Habitat | Area*$\left(\mathrm{m}^{2}\right)$ | Habitat Cover (\%) | No. of Species Observed | No. of Fish Observed (Mean No. Fish/100 m²) |  |  |  | Density |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total | Fish / $100 \mathrm{~m}^{2}$ | $\begin{gathered} \text { Species / } \\ 100 \mathrm{~m}^{2} \\ \hline \end{gathered}$ |
| Seagrass | 57326 | 79.62 | 81 | 8492 (14.8) | 3907 (6.8) | 1334 (2.32) | 13733 | 24 | 0.1 |
| Algal Plain | 9052 | 12.57 | 58 | 1662 (18.4) | 409 (4.5) | 1019 (11.25) | 3090 | 34.1 | 0.6 |
| Patch Reef | 2836 | 3.94 | 63 | 1019 (35.9) | 832 (29.3) | 349 (12.3) | 2200 | 77.5 | 2.2 |
| Sand | 2462 | 3.42 | 12 | 207 (8.4) | 47 (1.9) | 25 (1.01) | 279 | 11.3 | 0.5 |
| Rubble | 324 | 0.45 | 31 | 324 (100) | 44 (13.6) | 18 (5.55) | 386 | 119.1 | 9.5 |
| Total | 72000 |  |  | 11704 | 5239 | 2745 | 19688 |  |  |

*note: Area $=$ Total area of habitat surveyed in transects.

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Table 3. Number of Fish Observed and Fish Densities By Fish Size Groups and Northeast St. Croix Embayments; Number of Species per $100 \mathrm{~m}^{2}$ by Embayment; and Number of Fish Observed and Fish Densities by Fish Size Groups by Habitat and Northeast St. Croix Embavment (October 1998 to September 1999)

| $\overline{\text { Site }}$ | Bay Size | Survey | No. of | No. of Fish Observed (Fish/100m²) |  |  |  | Species$/ 100 \mathrm{~m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} (* 1) \\ \left(\mathrm{km}^{2}\right) \end{gathered}$ | $\begin{gathered} \text { Area(*2) } \\ \left(\mathrm{m}^{2}\right) \end{gathered}$ | Observed | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total |  |
| Cottongarden Bay | 0.41 | 24,000 | 54 | 1956 (8.1) | 1873 (7.8) | 642 (2.7) | 4471 (18.6) | 0.2 |
| Teague Bay | 0.76 | 24,000 | 66 | 4687 (19.5) | 1555 (6.5) | 1587 (6.61) | 7829 (32.6) | 0.2 |
| Yellowcliff Bay | 0.32 | 24,000 | 74 | 5061 (21.1) | 1811 (7.5) | 516 (2.15) | 7388 (30.8) | 0.3 |
| Total |  | 72,000 |  | 11704 (16.3) | 5239 (7.3) | 2745 (3.8) | 19688 (27.3) |  |


|  | Habitat Habitat <br> Area Surveyed <br> $\left({ }^{* 3)}\right.$ $\left({ }^{* 4)}\right.$ <br> $\left(\mathrm{km}^{2}\right)$ $\left(\mathrm{m}^{2}\right)$ |  | No. of Fish Observed (Fish/100 $\mathrm{m}^{2}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Overall |
| Cottongarden Bay |  |  |  |  |  |  |
| Seagrass | 0.37 | 22293 | 1329 (6.0) | 1419 (6.4) | 455 (2) | 3203 (14.4) |
| Algal Plain | 0.0 | 19 | 5 (26.3) | 0 | 1 (5.3) | 6 (31.6) |
| Patch Reef | 0.02 | 1144 | 512 (44.8) | 425 (37.2) | 182 (15.9) | 1119 (97.8) |
| Sand | 0.01 | 411 | 0 | 0 | 1 (0.2) | 1 (0.2) |
| Rubble | 0.01 | 133 | 110 (82.7) | 29 (21.8) | 3 (2.3) | 142 (106.8) |
| Teague Bay |  |  |  |  |  |  |
| Seagrass | 0.64 | 20548 | 4089 (19.9) | 1315 (6.4) | 559 (2.7) | 5963 (29.0) |
| Algal Plain | 0.08 | 2294 | 351 (15.3) | 111 (4.8) | . 947 (41.3) | 1409 (61.4) |
| Patch Reef | 0.02 | 266 | 167 (62.8) | 59 (22.2) | 59 (22.2) | 285 (107.1) |
| Sand | 0.02 | 879 | 75 (8.5) | 70 (8) | 22 (2.5) | 167 (19) |
| Rubble | 0 | 13 | 5 (38.5) | 0 | 0 | 5 (38.5) |


| Yellowcliff Bay |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Seagrass | 0.20 | 14614 | $2956(20.2)$ | $974(6.7)$ | $232(1.6)$ | $4162(28.5)$ |
| Algal Plain | 0.08 | 6180 | $1462(23.7)$ | $323(5.2)$ | $89(1.4)$ | $1874(30.3)$ |
| Patch Reef | 0.02 | 1360 | $281(20.7)$ | $446(32.8)$ | $171(12.6)$ | $898(66)$ |
| Sand | 0.01 | 1724 | $149(8.6)$ | $45(2.6)$ | $9(0.5)$ | $203(11.8)$ |
| Rubble | 0.01 | 122 | $213(174.6)$ | $23(18.9)$ | $15(12.3)$ | $251(205.7)$ |

*notes:

1. Area Estimated using a planimeter:
2. Survey area is the area surveyed by the benthic visual survey within each embayment.
3. Habitat Area = Bay Size estimated using a planimeter $x$ Habitat Percent Cover estimated from benthic visual transects.
4. Habitat Surveyed is the area of each habitat surveyed by the benthic visual survey within each embayment.

Table 4. Percent Similarity Values Between Habitats Sampled Based on Percent Similarity of Species Composition and Fish Densities From Pooled Monthly, Transect Site, and Northeast St. Croix Embayment Data (October 1998 to September 1999).

|  | Rubble | Sand | Algal Plain | Patch Reef | Seagrass |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Rubble | 11 | 37 | 45 | 44 |  |
| Sand |  | 15 | 10 | 15 |  |
| Algal Plain |  |  | 24 | 69 |  |
| Patch Reef |  |  |  | $\mathbf{5 3}$ |  |
| Seagrass |  |  |  |  |  |

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Table 5. Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Northeast St. Croix Embayments, Based on Pooled Monthly, Transect Site, and Embayment Data between October 1998 and September 1999.

\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline \& \multicolumn{6}{|l|}{No. of Fish in Each Fish Size Groups} \& \multicolumn{6}{|c|}{No. of Fish/100m ${ }^{2}$} <br>
\hline \multirow[t]{2}{*}{Species} \& \multicolumn{6}{|l|}{Family Acanthuridae} \& \& \& \& \& \& \multirow[b]{2}{*}{Total} <br>
\hline \& Habitat \& $<3 \mathrm{~cm}$ \& 65 cm \& $5-10 \mathrm{~cm}>$ \& 10 cm \& Total \& \multirow[t]{2}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{$<5 \mathrm{~cm}$} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& <br>
\hline \multirow{6}{*}{A. bahianus} \& Patch Reef \& \& $0 \quad 9$ \& 15 \& 53 \& 77 \& \& 0 \& 0.32 \& 0.53 \& 1.86 \& 2.71 <br>
\hline \& Rubble \& \& \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& <br>
\hline \& Sand \& \& $0 \quad 0$ \& 0 \& \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Algal Plaiu \& \& \& 0 \& 0 \& 2 \& \& \& 0.22 \& 0 \& \& 0.02 <br>
\hline \& Seagrass \& \& \& 4 \& 11 \& 40 \& \& 0 \& 0.04 \& 0.007 \& 0.02 \& . 07 <br>
\hline \& Total \& \& \& 19 \& 64 \& 119 \& \& \& \& \& \& <br>
\hline \multirow[t]{7}{*}{A.chirurgus} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{4}{|l|}{$<5 \mathrm{~cm} \quad 5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total} \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{$<5 \mathrm{~cm}$} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& $0 \quad 43$ \& 74 \& 76 \& 193 \& \& 0 \& 1.51 \& 2.61 \& 2.68 \& 6.8 <br>
\hline \& Rubble \& \& 0.36 \& 5 \& \& 51 \& \& \& 11.11 \& 1.54 \& 3.08 \& 15.7 <br>
\hline \& Sand \& \& 00 \& 0 \& 0 \& 0 \& \& 0 \& . \& 0 \& 0 \& <br>
\hline \& Algal Plain \& \& $0 \quad 8$ \& 37 \& 12 \& 12 \& \& 0 \& 0.08 \& 0.41 \& 0.13 \& 0.13 <br>
\hline \& Seagrass \& \& $0 \quad 143$ \& 55 \& 22 \& 220 \& \& 0 \& 0.25 \& 0.09 \& 0.04 \& 0.38 <br>
\hline \& Total \& \& 0.230 \& 171. \& 120 \& 476 \& \& \& \& \& \& <br>
\hline \multirow[t]{8}{*}{A coerulets} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$

0} \& \multicolumn{4}{|l|}{$55 \mathrm{~cm} \quad 5-10 \mathrm{~cm} \geqslant 10 \mathrm{~cm}$ Total} \& \multirow[t]{8}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{$<5 \mathrm{~cm}$} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& $0 \quad 25$ \& . 30 \& 14 \& 69 \& \& 0 \& 0.88 \& 1.05 \& 0.49 \& 2.43 <br>
\hline \& Rubble \& \& 02 \& 0 \& 0 \& 2 \& \& 0 \& . 62 \& 0 \& 0 \& 0.62 <br>
\hline \& Sand \& \& \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Algal Plain \& \& \& 0 \& \& 0 \& \& \& \& 0 \& \& 0 <br>
\hline \& Seagrass \& \& \& 4 \& 0 \& 8 \& \& \& 0.007 \& 0.007 \& 0 \& 0.01 <br>
\hline \& Total \& \& $0 \times 31$ \& 34 \& 14 \& 79 \& \& \& \& \& \& <br>
\hline \& \multicolumn{6}{|l|}{Family Haemulidae} \& \& \& \& \& \& <br>
\hline \multirow[t]{7}{*}{H. aurolineatum} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& $<5 \mathrm{~cm}$ \& \multicolumn{3}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total} \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$
$\because$
0} \& \multicolumn{2}{|r|}{$<5 \mathrm{~cm}$} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& 0.5 \& 21 \& 26 \& 52 \& \& 0 \& 0.17 \& 0.74 \& 0.92 \& 1.83 <br>
\hline \& Rubble \& \& \& 0 \& \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Sand \& \& 0 - 0 \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& <br>
\hline \& Algal Plain \& \& 8 - 30 \& \& 600 \& 648 \& \& 0.2 \& 0.33 \& 0 \& 6.62 \& 7.15 <br>
\hline \& Seagrass \& \& $0 \quad 386$ \& 145 \& 2 \& 533 \& \& 0 \& 0.67 \& 0.25 \& 0.35 \& 0.92 <br>
\hline \& Total \& \& 8 421 \& 166 \& 628 \& 1233 \& \& \& \& \& \& <br>
\hline \multirow[t]{7}{*}{H. carbonarium} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& < 5 cm \& \multicolumn{3}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total} \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{$<5 \mathrm{~cm}$} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& $0: 0$ \& 6 \& 7 \& 13 \& \& 0 \& 0 \& 0.21 \& 0.25 \& 0.46 <br>
\hline \& Rubble \& \& \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& <br>
\hline \& Sand \& \& 00 \& 0 \& 0 \& 0 \& \& 0 \& 0. \& 0 \& 0 \& 0 <br>
\hline \& Algal Plain \& \& $0 \quad 0$ \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Seagrass \& \& $0 \quad 0$ \& 6 \& 0 \& 6 \& \& 0 \& 0 \& 0.001 \& 0 \& 0.001 <br>
\hline \& Total \& \& $0 \quad 0$ \& 6 \& 7 \& \& \& \& \& \& \& <br>
\hline \multirow[t]{7}{*}{H. flavolineatum} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& $<5 \mathrm{~cm}$ \& \multicolumn{3}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total} \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{65cm} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& 088 \& 60 \& 27 \& 175 \& \& 0 \& 3.10 \& 2.11 \& 0.95 \& 6.17 <br>
\hline \& Rubble \& \& $0 \quad 0$ \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Sand \& \& $0 \quad 0$ \& 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Algal Plain \& \& $0 \quad 0$ \& 0 \& 100 \& 100 \& \& 0 \& 0 \& 0 \& 1.10 \& 1.10 <br>
\hline \& Seagrass \& \& 0.91 \& 300 \& 9 \& 400 \& \& 0 \& 0.16 \& 0.52 \& . 01 \& 0.69 <br>
\hline \& Total \& \& $0 \quad 179$ \& 360 \& 136 \& 675 \& \& \& \& \& \& <br>
\hline \multirow[t]{7}{*}{H. plumiert} \& Habitat \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& $<5 \mathrm{~cm}$ \& \multicolumn{3}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total} \& \multirow[t]{7}{*}{$<3 \mathrm{~cm}$} \& \multicolumn{2}{|r|}{< 5 cm} \& \multicolumn{2}{|l|}{$5-10 \mathrm{~cm}>10 \mathrm{~cm}$} \& Total <br>
\hline \& Patch Reef \& \& $0 \quad 16$ \& - 32 \& 50 \& 98 \& \& 0 \& 0.56 \& 1.13 \& 0.02 \& 0.03 <br>
\hline \& Rubble \& \& $0 \quad 0$ \& 0 - 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Sand \& \& $0 \quad 0$ \& 0 0 \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Algal Plain \& \& $0 \quad 0$ \& $0 \times$ \& 0 \& 0 \& \& 0 \& 0 \& 0 \& 0 \& 0 <br>
\hline \& Seagrass \& \& 0.66 \& -11 \& 1 \& 78 \& \& 0 \& 0.11 \& 0.02 \& 0.02 \& 0.14 <br>
\hline \& Total \& \& $0 \quad 82$ \& 2 - 43 \& 51 \& 176 \& \& \& \& \& \& <br>
\hline
\end{tabular}

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Table 5 (Continued). Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Northeast St. Croix Embayments, Based on Pooled Monthly, Transect Site, and Embayment Data between October 1998 and September 1999.


Table 5 (Continued). Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Northeast St. Croix Embayments, Based on Pooled Monthly, Transect Site, and Embayment Data between October 1998 and September 1999,

| No. of Fish in Each Size Group |  |  |  |  |  |  |  |  | No. of Fish/100m ${ }^{2}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species <br> S. taeniopterus | Family Scaridae |  | $<5 \mathrm{~cm}$ |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  |  | $<3 \mathrm{~cm}$ | < 5 cm |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ |  | Total | 0.08 |
|  | Habitat | 43 cm |  |  |  |  |  |  |  |  |  |  |
|  | Patch Reef |  | 0 | 165 |  |  |  | 62 | 2 |  | 229 |  | 0 | 5.81 |  | 2.18 | 0.07 |  |
|  | Rubble |  | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Sand |  | 0 |  | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain | 0 |  | 5 | 1 | 0 |  | 6 |  | 0 | 0.05 | 0.01 | 0 |  | 0.06 |
|  | Seagrass | 0 |  | 92 | 6 | 1 |  | 99 |  | 0 | 0.16 | 0.01 | 0.002 |  | 0.17 |
|  | Total |  |  | $0 \quad 262$ | 69 | 3 |  | 334 |  |  |  |  |  |  | 0 |
| S aurofrenatum | Habitat | 43 mm | $\leqslant \mathrm{cm}$ |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  |  | $<3 \mathrm{~cm}$ | $<5 \mathrm{~cm}$ |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  |  |
|  | Patch Reef |  | 00 | 2 | 10 | 14 |  | 26 |  | 0 | 0.07 | 0.35 | 0.49 |  | 0.91 |
|  | Rubble |  |  | 0 | 2 | 0 |  | 2 |  | 0 | 0 | 0.61 | 0 |  | 0.62 |
|  | Sand |  | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain |  | 0 | 0 | 0 | 3 |  | 3 |  | 0 | 0 | 0 | 0.003 |  | 0.003 |
|  | Seagrass |  |  | $0-2$ | 0 | 8 |  | 10 |  | 0 | 0.003 | 0 | 0.001 |  | 0.001 |
|  | Total |  | 0 |  | 12 | 25 |  | 41 |  |  |  |  |  |  |  |
| S. viride | Habitat | $<3 \mathrm{~cm}$ | $<5 \mathrm{~cm}$ |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  |  | <3cm | S5cm |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  |  |
|  | Patch Reef |  | 0 | 26 | 51 | 8 |  | 85 |  | 0 | 0.92 | 1.79 | 0.28 |  | 2.99 |
|  | Rubble |  | 0 | 0 | 1 | 0 |  | 1 |  | 0 | 0 | 0.36 | 0 |  | 0.31 |
|  | Sand |  | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain |  | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Seagrass |  | 0 | 0 | 0 | 0 |  | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Total |  | 0 | 26 | 52 | 8 |  | 86 |  |  |  |  |  |  |  |

Table 6. Total Number of Fish and Species Caught by Fish Trap and Percentage (Number of a Fish Species Caught/Total Number of Fish Caught in an Embayment) for Each Northeast St. Croix Embayment (120 trapdays/embavment) between October 1998 and Sepptember 1999 Based on Pooled Monthly Data

| FAMILY | SPECIES | Cottongard No of Fish | en Bay <br> Percent | Teague No. of Fish | Bay <br> Percent | Yellowcliff No. of Fish | fBay <br> Percent |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Acanthuridae | Acanthurus chirurgus | 10 | 1.44 | 18 | 2.37 | 7 | 1.60 |
|  | Acanthirus coeruleus | * | * | * | * | 6 | 1.37 |
| Aulostomidae | Aulostomus maculatus | 2 | 0.29 | 5 | 0.66 | 7 | 1.60 |
| Chaetodontidae | Chaetodon capistratus | 8 | 1.15 | 14 | 1.84 | 18 | 4.11 |
|  | Chaetodon striatus | 1 | 0.14 | * * | * | * | * |
| Gerreidae | Eucinostomus jonesi | * | * | 4 | 0.53 | * | * |
|  | Gerres cinereus | * | * | 1 | 0.13 | * | * |
| Haemulidae | Haemulon aurolineatum | * | * | 43 | 5.66 | 9 | 2.05 |
|  | Haemulon chrysargyreum | * | * | 32 | 4.21 | * | * |
|  | Haemulon flavolineatum | 3 | 0.43 | 239 | 31.45 | 24 | 5.48 |
|  | Haemulon plumieri | 10 | 1.44 | 27 | 3.55 | 16 | 3.65 |
| Holocentridae | Holocentrus adscensionis | 54 | 7.76 | 25 | 3.29 | 8 | 1.83 |
|  | Myripristis jacobus | 8 | 1.15 | 3 | 0.39 | 19 | 4.34 |
| Labridae | Halichoeres bivittatus | 3 | 0.43 | 3 | 0.39 | * | * |
| Lutianidae | Lutjanus analis | * | * | 5 | 0.66 | * | * |
|  | Lutjanus mahogani | 1 | 0.14 | 3 | 0.39 | 7. | 1.60 |
|  | Lutjanus synagris | * | * | 35 | 4.61 | 15 | 3.42 |
|  | Ocyurus chrysurus | 32 | 4.60 | 79 | 10.39 | 40 | 9.13 |
| Monacanthidae | Monacanthus ciliatus | 5 | 0.72 | 10 | 1.32 | * | * |
| Mullidae | Pseudupeneus maculatus | 269 | 38.65 | 146 | 19.21 | 209 | 47.72 |
| Muraenidae | Gymnothorax moringa | 6 | 0.86 | 2 | 0.26 | 1 | 0.23 |
| Pomacanthidae | Pomacanthus pani | * | * | 5 | 0.66 | * | * |
|  | Stegastes leucostictus | * | * | 1 | 0.13 | * | * |
| Scaridae | Sparisoma radians | 267 | 38.36 | 38 | 5.00 | 30 | 6.85 |
| Sciaenidae | Sciaenops oculatus | * | * | 2 | 0.26 | * | * |
| Serranidae | Alphestes afer | 16 | 2.30 | * | * | 2 | 0.46 |
|  | Epinephelus guttatus | * | * | 1 | 0.13 | 1 | 0.23 |
|  | Epinephelus striatus | * | * | 1 | 0.13 | 1 | 0.23 |
| Sparidae | Archosargus rhomboidalis | 1 | 0.14 | * | * | * | * |
| Tetradontidae | Canthigaster rostrata | * | * | 5 | 0.66 | 1 | 0.23 |
|  | Sphoeroides spengleri | * | * | 13 | 1.71 | 17 | 3.88 |
|  | Total Number of Fishes | 696 |  | 760 |  | 438 |  |
|  | Total Number of Species | 16 |  | 26 |  | 19 |  |



[^0]Table 7. Total Number of Fish and Species Caught by Seine Net and Percentage (Number of a Fish Species Caught/Total Number of Fish Caught in an Embayment) for Each Northeast St. Croix Embayment ( 24 seine hauls/embayment) between October 1998 and September 1999 Based on Pooled Monthly Data.

| FAMILY | SPECIES | Cottongarden Bay |  | Teague Bay |  | Yellowcliff Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish | Percent | No. of Fish | Percent | No. of Fish | Percent |
| Acanthuridae | Acanthurus chirurgus | * | * | 1 | 0.38 | * ${ }^{\text {* }}$ | * |
| Abbulidae | Albula vulpes | * | * | 2 | 0.76 | * | * |
| Belonidae | Tylosorus crocodrilus | 4 | 6.56 | 1 | 0.38 | 3 | 30 |
| Carangidae | Caranx hippos | * | * | 2 | 0.76 | * | * |
|  | Caranx latus | 10 | 16.39 | 74 | 28.14 | * | * |
|  | Selar crumenophtalmus | 5 | 8.20 | 1 | 0.38 | * | * |
|  | Selene vomer | * | * | 1 | 0.38 | * | * |
|  | Trachinotus falcatus | 23 | 37.70 | 5 | 1.90 | * | * |
| Chaetodontidae | Chaetodon capistratus | * | * | * | * | 1 | 10 |
| Clupeidae | Harengula humeralis | 1 | 1.64 | 14 | 5.32 | * | * |
|  | Harengula clupeola | 4 | 6.56 | 6 | 2.28 | * | * |
| Dactylopteridae | Dactylopterus volitans | * | * | 2 | 0.76 | * | * |
| Engraulidae | Anchoa unk. | * | * | 6 | 2.28 | * | * |
|  | Hypoatherina harringtonensis | * | * | 4 | 1.52 | * | * |
| Exocoetidae | Hemiramphus braciliencis | 1 | 1.64 | 3 | 1.14 | * | * |
| Gerreidae | Eucinostomus jonesi | 12 | 19.67 | 51 | 19.39 | 5 | 50 |
| Haemulidae | Haemulon flavolineatum | * | * | 4 | - 1.52 | * | * |
| Lutjanidae | Lutjamus analis | * | * | 2 | 0.76 | * | * |
|  | Lutjams apodus | * | * | 8 | 3.04 | * | * |
|  | Lutjamus mahogani | * | * | 7 | 2.66 | * | * |
|  | Lutjanus synagris | * | * | 1 | 0.38 | * | * |
|  | Ocyurus chrysurus | * | * | 27 | 10.27 | * | * |
| Monacanthidae | Monacanthus cilictus | * | * | 2 | 0.76 | * | * |
| Mullidae | Pseudupeneus maculatus | * | * | 3 | 1.14 | * | * |
| Pomacentridae | Abudefduf saxatilis | 1 | 1.64 | * | * | * | * |
|  | Microspathodon chrysurus | * | * | 1 | 0.38 | * | * |
|  | Stegastes unk. | * | * | 1 | 0.38 | * | * |
| Scaridae | Sparisoma radians | * | * | 3 | 1.14 | 1 | 10 |
| Sparidae | Calamus bajonado | * | * | 2 | 0.76 | * | * |
| Sphyraenidae | Sphyraena barracuda | * | * | 12 | 4.56 | * | * |
| Synodontidae | Synodontus unk. | * | * | 1 | 0.38 | * | * |
| Tetradontidae | Sphoeroides spengleri | * | * | 16 | 6.08 | * | * |
|  | Total Number of Fish | 61 |  | 263 |  | 10 |  |
|  | Total Number of Species | 9 |  | 29 |  | 4 |  |

Mean (Standard Error) Community Parameters of fish caught with seine nets during October 1998 to September 1999.

|  | Cottongarden <br> Bay | Teague Bay | Yellowcliff Bay |
| :--- | :---: | :---: | :---: |
| No. of fish/seine haul | $2.65(0.66)$ | $10.20(2.35)$ | $0.42(0.04)$ |
| No. of species/seine haul | $1.13(0.21)$ | $3.37(0.39)$ | $0.17(0.02)$ |

*not caught

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Table 8. Total Number of Fish and Species Observed, and Percentage (Number of a Fish Species Observed/Total Number of Fish Observed in an Embayment) for Each Southeast St. Croix Embavment (30 Transects Per Embayment) from Julv 2000 to September 2000 Based on Pooled Monthly and Transect Site Data.

| FAMILY | SPECIES | Turner Hole Bay |  | Robin Bay |  | Great Pond Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish |  | No. of Fish | Percent | No. of Fish |  |
| Acanthuridae | Acantharus bahianus | 16 | 0.98 | 4 | 1.29 | 7.00 | 0.54 |
|  | Acanthurus chirurgus | 189 | 11.61 | 7 | 2,27 | 12.00 | 0.93 |
|  | Acanthurus coeruleus | 15 | 0.92 | * | * | 10 | 0.78 |
| Apogonidae | Apogon unk. | 1 | 0.06 | 1 | 0.32 | * | * |
|  | Apogon stellatus | * | * | * | * | 2.00 | 0.16 |
| Aulostomidae | Aulostomus maculatus | 1 | 0.06 | * | * | * | * |
|  | Caranx ruber | 3 | 0.18 | 2 | 0.65 | 25.00 | 1.95 |
| Chaetodontidae | Chaetodon capistratus | 3 | 0.18 | 2 | 0.65 | 3.00 | 0.23 |
| Gerreidae | Gerres cinereus | 1 | 0.06 | 1 | 0.32 | * | * |
| Gobiidae | Coryphopterus glaucofraenum | 2 | 0.12 | * | * | 7.00 | 0.54 |
| Haemulidae | Haemulon unk. | 215 | 13.21 | 74 | 23.95 | 883.00 | 68.72 |
|  | Haemulon aurolineatum | 210 | 12.90 | * | * | 75.00 | 5.84 |
|  | Haemulon carbonarium | * | * | * | * | 1.00 | 0.08 |
|  | Haemulon chrysargyreum | * | * | * | * | 20 | 1.56 |
|  | Haemulon flavolineatum | 59 | 3.62 | 1 | 0.32 | 51.00 | 3.97 |
|  | Haemulon plumieri | 44 | 2.70 | * | * | * | * |
|  | Haemulon sciurum | * | * | * | * | 6.00 | 0.47 |
|  | Haemulon striatum | * | * | * | * | 7.00 | 0.54 |
| Holocentridae | Holocentrus unk. | 34 | 2.09 | 3 | 0.97 | * | * |
|  | Myripristis jacobus | 21 | 1.29 | 1 | 0.32 | * | * |
| Labridae | Bodiamus rufus | * | * | * | * | 1.00 | 0.08 |
|  | Doratonotus megalepsis | * | * | 1 | 0.32 | * | * |
|  | Halichoeres unk. | * | * | * | * | * | * |
|  | Halichoeres bivittatus | 207 | 12.71 | 93 | 30.10 | 104.00 | 8.09 |
|  | Halichoeres garnoti | 9 | 0.55 | * | * | * | * |
|  | Halichoeres poeyi | 4 | 0.25 | * | * | * | * |
|  | Hemipteronotus martinicensis | * | * | * | * | 2.00 | 0.16 |
|  | Hemipteronotus splendens | * | * | * | * | 2.00 | 0.16 |
|  | Thalassoma bifasciatum | 241 | 14.80 | 14 | 4.53 | 12.00 | 0.93 |
| Lutjanidae | Lutjanus mahogani | 2 | 0.12 | * | * | 1.00 | 0.08 |
|  | Lutjamus synagris | 1 | 0.06 | * | * | 3.00 | 0.23 |
|  | Ocyurus chrysurus | 22 | 1.35 | * | * | 2.00 | 0.16 |
| Monacanthidae | Monacanthus ciliatus | * | * | 3 | 0.97 | 1.00 | 0.08 |
| Mullidae | Pseudupeneus maculatus | 8 | 0.49 | 2 | 0.65 | * | * |
| Ophichthidae | Myrichthys ocellatus | * | * | 1 | 0.32 | * | * |
| Ostraciidae | Lactophrys triqueter | * | * | 1 | 0.32 | * | * |
| Pomacentridae | Chromys cyanea | 2 | 0.12 | * | * | * | * |
|  | Microspathodon chrysurus | 7 | 0.43 | 1 | 0.32 | ${ }^{*}$ | ${ }^{*}$ |
|  | Stegastes fuscus | 19 | 1.17 | 7 | 2.27 | 2.00 | 0.16 |
|  | Stegastes leucostictus | 90 | 5.53 | 2 | 0.65 | 7.00 | 0.54 |
|  | Stegastes partitus | 18 | 1.11 | * | * | 3.00 | 0.23 |
|  | Stegastes planifrons | 4 | 0.25 | * | * | * | * |
|  | Stegastes variabilis | * | * | * | * * | 1.00 | 0.08 |
| Scaridae | Cryptotomus roseus | * | * | 2 | 0.65 | 1.00 | 0.08 |

Table 8 (Continued). Total Number of Fish and Species Observed, and Percentage (Number of a Fish Species Observed/Total Number of Fish Observed in an Embayment) for Each Southeast St, Croix Embayment ( 30 Transects Per Embayment) from July 2000 to September 2000 Based on Pooled Monthly and Transect Site Data.

| FAMLIY | SPECIES | Turner Hole Bay |  | Robin Bay |  | Great Pond Bay |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. of Fish | Percent | No. of Fish | Percent | No. of Fish | Percent |
| Scaridae | Scarus iserti | 1 | 0.06 | * | * | 7.00 | 0.54 |
|  | Scarus taeniopterus | 17 | 1.04 | * | * | 8.00 | 0.62 |
|  | Sparisoma cuirofrenatum | 12 | 0.74 | 1 | 0.32 | * | * |
|  | Sparisoma radians | 104 | 6.39 | 76 | 24.60 | 5.00 | 0.39 |
|  | Sparisoma viride | 16 | 0.98 | 5 | 1.62 | 7.00 | 0.54 |
|  | Sparisoma chrysopterum | 7 | 0.43 | * | * | * | * |
| Sciaenidae | Equetus acuminatus | 5 | 0.31 | * | * | * | * |
| Serranidae | Epinephelus fulvis | 6 | 0.37 | * | * | * | * |
|  | Epinephelus guttatus | 11 | 0.68 | * | * | 1.00 | 0.08 |
|  | Serramus tabacarius | 1 | 0.06 | * | * | * | * |
|  | Serranus tigrinus | * | * | * | * | 3.00 | 0.23 |
| Tetradontidae | Canthigaster rostrata | * | * | , | 0.32 | 2.00 | 0.16 |
|  | Sphoeroides spengleri | * | * | 3 | 0.97 | 1.00 | 0.08 |
|  | Total Number of Fish | 1628 |  | 309 |  | 1285 |  |
|  | Total Number of Species | 39 |  | 25 |  | 36 |  |

Mean (standard error) community parameters of fish censused with visual transects during
July 2000 to September 2000. All months were pooled

|  | Turner Hole Bay | Robin Bay | Great Pond Bay |
| :--- | :---: | :---: | :---: |
| No. of fish $/ 100 \mathrm{~m}^{2}$ | $26.6(4.40)$ | $5.90(1.49)$ | $19.45(5.21)$ |
| No. of species $/ 100 \mathrm{~m}^{2}$ | $3.75(0.52)$ | $1.55(0.29)$ | $1.70(0.27)$ |
| Diversity Index | $0.76(0.09)$ | $0.30(0.07)$ | $0.28(0.07)$ |
| Evenness | $0.62(0.05)$ | $0.63(0.06)$ | $0.69(0.05)$ |

*not observed

Table 9. Number of Fish Observed and Densities By Size Groups and By Habitat Based on Pooled Monthly, Transect Site; and Southeast St. Croix Embayment Data (July to September 2000).

| Habitat | Area*$\left(\mathrm{m}^{2}\right)$ | Habitat <br> Cover <br> ( $\mathrm{m}^{2}$ ) | No, of <br> Species Observed | No. of Fish Observed (Mean No. Fish/100m ${ }^{2}$ ) |  |  |  | Density |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total | Fish <br> $100 \mathrm{~m}^{2}$ | Species $100 \mathrm{~m}^{2}$ |
| Seagrass | 10126 | 56.26 | 23 | 450 (4.4) | 151 (1.5) | 35 (0.34) | 636 | 6.3 | 0.2 |
| Sand | 6131 | 34.06 | 9 | 199 (3.2) | 0 | 29 (0.47) | 228 | 3.7 | 0.1 |
| Algal Plain | 1164 | 6.47 | 19 | 1002 (86.1) | 42 (3.6) | 0 | 1044 | 89.6 | 0.2 |
| Patch Reef | 358 | 1.99 | 41 | $697194.7)$ | 378 (105.6) | 155 (43.3) | 1230 | 343.5 | 11.5 |
| Rubble | 221 | 1.23 | 14 | 39 (17.6) | 62 (28) | 12 (5.4) | 113 | 51.1 | 6.3 |
| Total | 18000 |  |  | 2387 | 633 | 231 | 3251 |  |  |

[^1]Table 10. Number of Fish Observed and Fish Densities By Fish Size Groups and Southeast St. Croix Embayments; Number of Species per $100 \mathrm{~m}^{2}$ by Embayment; and Number of Fish Observed and Fish Densities by Fish Size Groups by Habitat and Southeast St. Croix Embavment (July to September 2000).

| Site | $\begin{gathered} \text { Bay Size } \\ \left({ }^{*} 1\right) \\ \left(\mathrm{km}^{2}\right) \\ \hline \end{gathered}$ | $\begin{gathered} \text { Survey } \\ \text { Area }(* 2) \\ \left(\mathrm{m}^{2}\right) \\ \hline \end{gathered}$ | No. of Species Observed | No. of Fish Observed (Fish/100m ${ }^{\text {2 }}$ ) |  |  |  | Species$/ 100 \mathrm{~m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| Turner Hole Bay | 0.37 | 6,000 | 39 | 998 (16.6) | 463 (7.7) | 172 (2.86) | 1633 (27.2) | 0.6 |
| Robin Bay | 0.64 | 6,000 | 25 | 177 (2.9) | 104 (1.7) | 29 (0.5) | 310 (5.2) | 0.4 |
| Great Pond Bay | 2.30 | 6,000 | 36 | 1212 (20.2) | 66 (1.1) | 30 (0.5) | 1308 (21.8) | 0.6 |
| Total |  | 18,000 |  | 2387 (13.3) | 633 (3.5) | 231 (1.3) | 3251 (18.1) |  |


| Total | Habitat Area (*3) ( $\mathrm{km}^{2}$ ) | Habitat Surveyed$\begin{aligned} & \left({ }^{*} 4\right) \\ & \left(\mathrm{m}^{2}\right) \end{aligned}$ | No. of Fish Observed (Fish/100 m${ }^{\text {2 }}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
|  |  |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Overall |
| Turner Hole Bay |  |  |  |  |  |  |
| Seagrass | 0.24 | 3,938 | 230 (5.8) | 63 (1.6) | 30 (0.8) | 323 (8.2) |
| Algal Plain | 0.02 | 300 | 163 (54.3) | 7 (2.3) | 0 | 170 (56.7) |
| Patch Reef | 0.02 | 288 | 564 (195.6) | $331(114.9)$ | 127 (44.1) | 1022 (354.9) |
| Sand | 0.07 | 1,253 | 2 (0.2) | 0 | 3 (0.2) | 5 (0.4) |
| Rubble | 0.01 | 221 | 39 (17.6) | 62 (28.1) | 12 (5.4) | 113 (51.1) |
| Robin Bay |  |  |  |  |  |  |
| Seagrass | 0.58 | 5,521 | 121 (2.2) | 87 (1.6) | 5 (0.1) | 213 (3.9) |
| Algal Plain | 0.01 | 22 | 10 (45.5) | 0 | 0 | 10 (45.5) |
| Patch Reef | 0.01 | 29 | 46 (158.6) | 17 (58.6) | 23 (79.3) | 86 (296.6) |
| Sand | - 0.04 | 428 | 0 | 0 | 1 (0.2) | 1 (0.2) |
| Rubble | 0.00 | 0 | 0 | 0 | 0 | 0 |
| Great Pond Bay |  |  |  |  |  |  |
| Seagrass | 0.23 | 598 | 99 (16.6) | 1 (0.2) | 0 | 100 (16.7) |
| Algal Plain | 0.32 | 842 | 829 (98.5) | 35 (4.2) | 0 | 864 (102.6) |
| Patch Reef | 0.02 | 41 | 87 (212.2) | 30 (73.2) | 5 (12.2) | 122 (297.6) |
| Sand | 1.73 | 4,519 | 197 (4.4) | 0 | 25 (0.6) | 222 (4.9) |
| Rubble | 0.00 | 0 | 0 | 0 | 0 | 0 |

*notes:

1. Area Estimated using a planimeter.
2. Survey area is the area surveyed by the benthic visual survey within each embayment.
3. Habitat Area $=$ Bay Size estimated using a planimeter $\times$ Habitat Percent Cover estimated from benthic visual transects.
4. Habitat Surveyed is the area of each habitat surveyed by the benthic visual survey within each embayment.

Table 11. Percent Similarity Values Between Habitats Sampled Based on Percent Similarity of Species Composition and Fish Densities From Pooled Monthly, Transect Site, and Southeast St. Croix Embayment Data (July to September 2000)

|  | Rubble | Sand | Algal Plain | Patch Reef |
| :--- | :---: | :---: | :---: | :---: |
| Rubble | 3 | 8.8 | Seagrass |  |
| Sand |  | $\mathbf{8 2}$ | 11 | 48 |
| Algal Plain |  | 32 | 42 |  |
| Patch Reef |  |  | 48 |  |
| Seagrass |  |  | 26 |  |

Table 12. Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Southeast St. Croix Embayments, Based on Pooled Monthly, Transect Site, and Embayment Data between July and September 2000


Table 12 (Continued). Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Southeast St. Croix Embayments, Based on Pooled Monthly, Transect Site, and Embayment Data between July and Sentember 2000


Table 12. (Continued). Number of Fish per Size Group and Fish Densities of Economically Important Species By Habitat for Southeast St. Croix Embavments, Based on Pooled Monthly, Transect Site, and Embayment Data between July and September 2000.

|  | No. of Fish in Each Fish Size Group |  |  |  |  |  |  | No. of Fish / $100 \mathrm{~m}^{2}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species <br> L. apodus | Family Lutjanidae |  |  |  | $5-10 \mathrm{~cm}>10 \mathrm{~cm}$ Total |  |  | $<3 \mathrm{~cm}$ | 65 cm |  | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total |  |
|  | Habitat | $\leqslant 3 \mathrm{~cm}$ | 45 cm |  |  |  |  |  |  |  |  |  |  |  |
|  | Patch Reef | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 |  |  |  | 0 |
|  | Rubble | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Sand | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain |  |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Seagrass | 0 |  | 0 | 1 | 0 |  |  | 0 | - 0 | 0.09 | 0 |  | 0.09 |
|  | Total | 0 | $\checkmark$ | 0 | 1 | 0 | -1 |  |  |  |  |  |  |  |
|  | Family Scari | dae |  |  |  |  |  |  |  |  |  |  |  |  |
| S. iserti | Habitat | $<3 \mathrm{~cm}$ | . 55 cm |  | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total | 43 cm |  | $\leqslant 5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ |  |  |
|  | Patch Reef | 0 |  | 1 | 0 | 0 | 1 |  | 0 | 0.28 | 0 | 0 |  | 0.28 |
|  | Rubble | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Sand | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plaiu | 0 |  | 3 | 4 | 0 | 7 |  | 0 | 0.26 | 0.34 | 0 |  | 0.6 |
|  | Seagrass | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Total | 0 |  | 4 | 4 | 0 | 8 |  |  |  |  |  |  |  |
| S. taeniopterus | Habitat | $<3 \mathrm{~cm}$ | $<5 \mathrm{~cm}$ |  | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total | 43 cm |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ |  |  |
|  | Patch Reef | 0 |  | 6 | 0 | 0 | 16 |  | 0 | 4.47 | 0 | 0 |  | 4.47 |
|  | Rubble | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Sand | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain | 0 |  | 0 | 0 | 8 | 8 |  | 0 | 0 | 0 | 0.69 |  | 0.69 |
|  | Seagrass | 0 |  | 1 | 0 | 0 | 1 |  | 0 | 0.09 | 0 | 0 |  | 0.09 |
|  | Total | 0 | 17 | 7 | 0 | 8 | 25 |  |  |  |  |  |  |  |
| S. aurofrenatum | Habitat | 3 cm | < 5 cm |  | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ | Total | $<3 \mathrm{~cm}$ |  | < 5 cm | $5-10 \mathrm{~cm}$ | $>10 \mathrm{~cm}$ |  |  |
|  | Patch Reef. | 0 |  | 5 |  | 8 | 13 |  | 0 | 1.39 | 0 | - 2,23 |  | 3.63 |
|  | Rubble | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Sand | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Seagrass | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Total |  |  | 5 |  | 8 | 13 |  |  |  |  |  |  |  |
| S. viride | Habitat | $<3 \mathrm{~cm}$ | 45 cm |  | $5-10 \mathrm{~cm}>$ | $>10 \mathrm{~cm}$ | Total | 63 cm |  | $<5 \mathrm{~cm}$ | $5-10 \mathrm{~cm}$ | >10cm |  |  |
|  | Patch Reef | 0 |  | 12 | 9.00 | 4 | 25 |  | 0 | 3.35 | 2.51 | 1.12 |  | 6.98 |
|  | Rubble | 0 | 0 | 2 |  | 1 | 3 |  | 0 | 0,9 | 0 | 0.45 |  | 1.36 |
|  | Sand | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Algal Plain | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  | 0 |
|  | Seagrass | 0 |  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | - 0 |  | 0 |
|  | Total | 0 | 14 | 14 | 9 | 5 | 28 |  |  |  |  |  |  |  |

Table 13. Total Number of Fish and Species Caught by Fish Trap and Percentage (Number of a Fish Species Caught/Total Number of Fish Caught in an Embayment) for Each Southeast St. Croix Embayment (30 trapdays/embayment) between July and September 2000 Based on Pooled Monthly Data.

| FAMILY | SPECIES | Turner Hole Bay No. of Percent |  | Robi <br> No. of Fish | Bay <br> Percent | Great $P$ No. of Fish | Pond Bay Percent |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Acanthuridae | Acanthurus chirurgus | 13 | 19.40 | * | * | 4 | 33.33 |
| Aulostomidae | Aulostomus maculatus | * | * | * | * |  | * * |
| Bothidae | Paralichthys spp | * | * | 1 | 2.63 | * | * |
| Chaetodontidae | Chaetodon capistratus | 5 | 7.46 | * | * | 3 | 25.00 |
| Haemulidae | Haemulon flavolineatum | 4 | 5.97 | 3 | 7.89 | 1 | 8.33 |
|  | Haemulon plumieri | 1. | 1.49 | * | * | * | * |
| Holocentridae | Holocentrus unk. | 3 | 4.48 | 5 | 13.16 | 1 | 8.33 |
|  | Myripristis jacobus | * | * | 2 | 5.26 | * | * |
| Labridae | Halichoeres bivitictus | 1 | 1.49 | * | * | * | * |
| Lutjanidae | Lutjanus synagris | * | * | * | * | 3 | 25.00 |
|  | Ocyurus chrysurus | 10 | 14.93 | * | * | * | * |
| Monacanthidae | Monacanthus ciliatus | * | * | 1 | 2.63 | * | * |
| Mullidae | Pseudupeneus maculatus | 7 | 10.45 | 2 | 5.26 | * | * |
| Pomacanthidae | Pomacanthus paru | 1 | 1.49 | * | * | * | * |
| Scaridae | Sparisoma radians | 15 | 22.39 | 24 | 63.16 | * | * |
| Serranidae | Alphestes afer | 2 | 2.99 | * | * | * | * |
|  | Epinephelus guttatus | 1 | 1.49 | * | * | * | * * |
| Sparidae | Calamus bajonado | 4 | 5.97 | * | * | * | * * |
|  | Total Number of Fishes | 67.00 |  | 38 |  | 12 |  |
|  | Total Number of Species | 13.00 |  | 7 |  | 5 |  |
|  | Mean (Standard Error) Community Parameters of fish caught with traps during July 2000 to September 2000. All months were pooled. |  |  |  |  |  |  |
|  |  | Turner Hole Bay |  | Robin Bay |  | Great Pond Bay |  |
|  | No. of fish/trap-day | 1.53 (0.79) |  | 2.30 (0.79) |  | 0.45 (0.22) |  |
|  | No. of species/trap-day | 1.00 (0.27) |  | 0.43 (0.12) |  | 0.33 (0.15) |  |

*not caught

Table 14. Total Number of Fish and Species Caught by Seine Net and Percentage (Number of a Fish Species Caught/Total Number of Fish Caught in an Embayment) for Each Southeast St. Croix Embayment (6 seine hauls/embayment) between July and September 2000 Based on Pooled Monthly Data.

| FAMILY | SPECIES |  | ole Bay Percent | No. of <br> Fish | in Bay Percent |  | ond Bay Percent |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Acanthuridae | Acanthurus bahianus | * | * | * | * | 1 | 5 |
| Carangidae | Trachinotus falcatus | 7 | 78 | 5 | 50 | * | * |
|  | Trachinotus goodei | 1 | 11 | * | * | * | * |
| Clupeidae | Jenkinsia lamprotaenia |  | * | * | * | 7 | 35 |
| Gerreidae | Eucinostomus jonesi | * | * | 2 | 20 | 2 | 10 |
|  | Gerres cinereus | * | * | * | * | 1 | 5 |
| Lutjanidae | Lutjomus apodus | * | * | 1 | 10 | 6 | 30 |
| Monacanthidae | Monacanthus ciliatus | * | * | 1 | 10 | * | * |
| Ostraciidae | Lactophrys polygonia | 1 | 11 | * | * | * | * |
| Sciaenidae | Umbrina coroides | * | * | 1 | 10 | * | * |
| Sphyraenidae | Sphyraena barracuda | * | * | * | * | 2 | 10 |
| Tetradontidae | Sphoeroides spengleri | * | * | * | * | 1 | 5 |
|  | Total Number of Fish | 9 |  | 10 |  | 20 |  |
|  | Total Number of Species | 3 |  | 5 |  | 7 |  |

Mean (Standard Error) Community Parameters of fish caught with seine nets during July 2000 to September 2000. All months were pooled.

|  | Turner Hole <br> Bay | Robin Bay | Great Pond Bay |
| :--- | :---: | :---: | :---: |
|  | $1.83(0.91)$ | $1.67(0.95)$ | $3.50(1.91)$ |
| No. of fish/seine haul | $1.33(0.61)$ | $0.83(0.40)$ | $1.33(0.61)$ |

$*_{\text {not caught }}$

Figure 1. Location of the three northeast St. Croix (U. S. V.I.) embayments surveyed (Cottongarden Bay, Teague Bay, and Yellowcliff Bay)


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Figure 2. Location of the three southeast St. Croix (U.S.V.L) embayments surveyed
(Turner Hold Bay, Robin Bay, and Great Pond Bay)


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Figure 3. Substrate Percentage Composition for Northeast St. Croix Embayments based on Pooled Month and Transect Site Data
(Total Area of Each Embayment Surveyed $=24,000 \mathrm{~m}^{2}$ ).


Figure 4a. Mean Monthly Fish Densities For Each Northeast St. Croix Embayment based on Pooled Transect Site Data.


Figure 5a. Monthly Shannon Wiener Diversity Index For Each Northeast St. Croix Embayment based on Pooled Transect Site Data.


Figure 4b. Mean Monthly Density of Species for Each Northeast St. Croix Embayment based on Pooled Transect Site Data.


Figure 5b. Monthly Pielou Evenness Index For Each Northeast St. Croix Embayment Based on Pooled Transect Site Data.


Legend: © $\mathbf{m}$ Cottongarden Bay Teague Bay Yelloweliff Bay

Figure 6a. Mean Fish Densities For Northeast St. Creix Embayment By Habitat Based on Pooled Monthly and Traisect Site Data.


Figure 7a. Diversity Indexes (H') For Northeast St. Croix Embayments By Habitat Based on Pooled Monthly and Transect Site Data.


Figurc 6b. Mean Species Densities For Northeast St. Croix Embayments By Habitat Based on Pooled Monthly and Transect Site Data.


Figure 7b. Evenness Index ( $\mathrm{J}^{\prime}$ ) For Northeast St. Croix Embayments By Habitat Based on Pooled Monthly and Transect Site Data,


Figure 8a. Mean Monthly Haemulon unk. Densities for Northeast St. Croix Embayments Based on Pooled Transect Site and Size Group Data.


Figure 9a. Mean Monthly Halichoeres bivittatus Densities for Northeast St. Croix Embayments Based on Pooied Transect Site and Size Group Data.


Figure 8b. Mean Monthly Haemulon unk. Recruit ( $<5 \mathrm{~cm}$ TL) Densities for Northeast St. Croix Embayments Based on Pooled Transect Site Data.


Figure 9b. Mean Monthly Halichoeres bivittatus Recruit ( $<5 \mathrm{~cm} \mathrm{TL}$ ) Densities for Northeast St. Croix Embayments Based on Pooled Transect Site Data.


Legend: Cottongarden Bay Teague Bay $\leftrightarrow$ Yellowcliff Bay

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Figure 10a. Mean Monthly Sparisoma radians Densities for Noitheast St. Croix Embayments Based on Pooled Transect Site and Size


Figure 10b. Mean Monthly Sparisoma radians Recruit ( $<S \mathrm{~cm}$ TL) Densities for Northeast St. Croix Embayments Based on Pooled Transect Site Data.


Figure 11b. Mean Monthly Ocyurus chrysurus Recruit ( $\langle 5 \mathrm{~cm}$ TL) Densities for Northeast St. Croix Embayments Based on Pooled Transect Site Data.


Figure 11a. Mean Monthly Ocyurus chrysurus Densities for Northeast
St. Croix Embayments Based on Pooled Transect Site and Size


Legend: Cottongarden Bay Teague Bay Yellowcliff Bay

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Figure 12a. Mean Number of Fish Caught Per Trap-Day in Northeastem St. Croix Embayments by Month (Pooled Trap-day Data).


Figure 12b. Mean Number of Species Caught Per Trap-day in Northeast


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Figure 13a. Monthly Mean Number of Haenndon flavolineatum Caught
Per Trap-Day for Northeast St. Croix Embayments Based on Pooled Trap-day Data


Figure 13c. Monthly Mean Number of Pseuchupenets maculatus Caught Per Trap-Day for Northeast St. Croix Embayments Based on Pooled


Figure 13b. Monthly Mean Number of Sparisoma radians Caught Per Trap-Day for Northeast St. Croix Embayments Based on Pooled



Legend: ©Cottongarden Bay Teague Bay $\uparrow$ Yellowcliff Bay

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Figure 14a. Length Frequency Distribution of Haemulon flavolineatum Caught by Traps in Cottongarden Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}-109$ ).


Mean Size at Sexual Maturity 160 mm TL (Billings and Mumro 1983)

Figure 14b. Length Frequency Distribution of Haemulon flavolineatum Caught by Traps in Teague Bay Based on Pooled Trap-day and Monthly Data $(\mathrm{N}=23)$.


Mean Size at Sexual Maturity 160 mm TL (Billings and Munro 1983)

Figure 14c. Length Frequency Distribution of Haemulon flavolineatum
Caught by Traps in Yellowcliff Bay Based on Pooled Trap-day and
Monthly Data $(\mathbb{N}-4)$.


Mean Size at Sexual Maturity 160 mm TL (Billings and Munro 1983)

Figure 15a. Length Frequency Distribution of Sparisoma radians Caught by Traps in Cottongarden Bay Based on Pooled Trap-day and Monthly Data $(\mathrm{N}=150)$.


Mean Size at Sexual Maturity = No Information Available

Figure 15b. Length Frequency Distribution of Sparisoma radians
Caught by Traps in Teague Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=36$ ).


Mean Size at Sexual Maturity = No Information Available

Figure 15c. Length Frequency Distribution of Sparisoma radians
Caught by Traps in Yelloweliff Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=33$ ).


Mean Size at Sexual Maturity = No Information Available

Figure 16a. Length Frequency Distribution of Pseudupeneus maculatus
Caught by Traps in Cottongarden Bay Based on Pooled Trap-day and Monthly Data $(\mathbb{N}=1.45$ ).


Mean Size at Sexual Maturity $=180 \mathrm{~mm}$ TL (Garcia-Cagide et al 1994)

Figure 16b. Length Frequency Distribution of Pseudupeneus maculatus
Caught by Traps in Teague Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=126$ ).


Mean Size at Sexual Maturity $=180 \mathrm{~mm}$ TL (Garcia-Cagide et al 1994)

Figure 160. Length Frequency Distribution of Pseudupeneus maculatus Caught by Traps in Yellowcliff Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=162$ ).


Mean Size at Sexual Maturity $=180 \mathrm{~mm}$ TL (Garcia-Cagide et al 1994)

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Figure 17a. Length Frequency Distribution of Ocyurus chrysurus Caught by Traps in Cottongarden Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=36$ ).


Mean Size at Sexual Maturity $=240 \mathrm{~mm}$ TL (Claro 1983)

Figure 17b. Length Frequency Distribution of Ocyurus chrysurus
Caught by Traps in Teague Bay Based on Pooled Trap-day and
Monthly Data ( $\mathrm{N}=59$ ).


Mean Size at Sexual Maturity $=240 \mathrm{~mm} \mathrm{TL}$ (Claro 1983)

Figure 17c. Length Frequency Distribution of Ocyurus chrysurus Caught by Traps in Yellowcliff Bay Based on Pooled Trap-day and Monthly Data ( $\mathrm{N}=25$ ).


Mean Size at Sexual Maturity $=240 \mathrm{~mm}$ TL (Claro 1983)

Figure 18a. Mean Monthly Number of Fish Caught Per Seine-Haul for
Northeast St. Croix Embayments Based on Pooled Seine-Haul Data


Figure 18b. Mean Monthly Number of Species Caught Per Seine-Hauls for Northeast St. Croix Embayments Based on Pooled Seine-Haul Data.


Legend: Cottongarden Bay Teague Bay © Yellowcliff Bay

Figure 19a. Monthly Mean Number of Caranx latus Caught per Seine
Net-Haul for Northeast St. Croix Embayments Based on Pooled
Seine Net-Haul Data.


Figure 19b. Monthly Mean Number of Eucinostomus jonesi Caught per Seine Net-Haul for Northeast St. Croix Embayments Based on Pooled Seine Net-Haul Data.


Figure 19c. Monthly Mean Number of Trachinotus falcatus Caught per Seine Net-Haul for Northeast St. Croix Embayments Based on Pooled Seine Net-Haul Data.


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Figure 20a. Length Frequency Distribution of Caranx latus Caught by Seine Net-Hauls in Cottongarden Bay Based on Pooled Monthly


Mean Size at Sexual Maturity $=370 \mathrm{~mm}$ TL (Thomson and Murro 1974)

Figure 20b. Length Frequency Distribution of Caranx latus Caught by Seine Net-Hauls in Teague Bay Based on Pooled Monthly and Seine Net-Haul Data( $\mathrm{N}=30$ ).


Mean Size at Sexual Maturity $=370 \mathrm{~mm}$ TL (Thomson and Munro 1974)

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Figure 20c. Length Frequency Distribution of Eucinostomus jonensi
Caught by Seine Net-Hauls in Cottongarden Bay Based on Pooled
Monthly and Seine Net-Haul Data ( $\mathrm{N}=9$ ).


Mean Size at Sexual Maturity $=$ No Information Available

Figure 20d. Length Frequency Distribution of Eucinostomus jonensi
Caught by Seine Net-Hauls in Teague Bay Based on Pooled
Monthly and Seine Net-Haul Data $(\mathrm{N}=32)$.


Mean Size at Sexual Maturity $=$ No Information Available

Figure 20e. Length Frequency Distribution of Eucinostomus jonensi
Caught by Seine Net-Hauls in Yellowcliff Bay Based on Pooled
Monthly and Seine Net-Haul Data ( $\mathrm{N}=11$ )


Mean Size at Sexual Maturity $=$ No Information Available

Figure 20f. Length Frequency Distribution of Trachinotus falcalus Caught by Seine Net-Hauls in Cottongarden Bay Based on Pooled
Monthly and Seine Net-Haul Data ( $\mathrm{N}-15$ ).


Mean Size at Sexual Maturity $\mathbf{- 2 5 0} \mathrm{mm}$ TL (García-Cagide et al 1994)
Cited in FishBase Ver. 1.2 (1995)

Figure 20g. Length Frequency Distribution of Trachinotus falcatus
Caught by Seine Net-Hauls in Teague Bay Based on Pooled


Mean Size at Sexual Maturity $=250 \mathrm{~mm}$ TL (García-Cagide et al 1994)

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Figure 21. Substrate Percentage Composition for Southeast
St. Croix Embayments Based on Pooled Month and Transect Site Data
(Total Area of Each Bay Covered on Transects $=6,000 \mathrm{~m}^{2}$ ).


Figure 22a Mean Fish Densities by Habitat for Each Southeast
St. Croix Embayment Based on Pooled Monthly and Transect Site Data,


Figure 23a. Shannon Wiener Diversity Indexes ( $\mathrm{H}^{\prime}$ ) by Habitat for Each Southeast St. Croix Embayment Based on Pooled Monthly


Figure 22b. Mean Species Density by Habitat for Each Southeast St. Croix Embayment Based on Pooled Monthly and Transect Site Data.


Figure 23b. Evenness Index (J) by Habitat for Each Southeast St. Croix Embayment Based on Pooled Monthily and Transect


Legend:
\& Seagrass $\boxminus$ Patch Reef $\boxminus$ Algal Plain $\because$ Sand $\square$ Rubble

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Figure 24a. Length Frequency Distribution of Acanthurus chirurgus Caugth by Traps in Southeast St. Croix Embayments


Mean Size at Sexual Maturity $=140 \mathrm{~mm} \mathrm{TL}$ (Garcia-Cagide et al 1994)

Figure 24c, Length Frequency Distribation of Pseudupeneus maculatus Caugth by Traps in Southeast St. Croix Embayments Based on Pooled Trap-haul, Month, and Embayment Data ( $\mathrm{N}=9$ ).


Mean Size at Sexual Maturity $=180 \mathrm{~mm}$ TL (García-Cagide et al 1994)

Figure 24b. Length Frequency Distribution of Sparisoma radians Caugth by Traps in Southeast St. Croix Embayments


Mean Size at Sexual Maturity - No Information Available

Figuse 24d. Length Frequency Distribution of Ocyurus chrysurus Caugth by Traps in Southeast St. Croix Embayments
Based on Pooled Trap-haul, Month, and Embayment Data ( $N=10$ ).


Mean Size at Sexual Maturity $=240 \mathrm{~mm}$ TL (Claro 1983)

Figure 25a. Length Frequency Distribution of Trachinotus
falcatus Caught by Seine Net-Hauls in Southeast St. Croix
Embayments Based on Pooled Seine-haul, Month, and Embayment
Data ( $\mathrm{N}-12$ ).


Mean Size at Sexual Maturity $=250 \mathrm{~mm}$ TL (Garcia et al 1994)

Figure 25b. Length Frequency Distribution of Lutjanus
apodus Caught by Seine Net-Hauls in Southeast St. Croix
Embayments Based on Pooled Seine-haul, Month and Embayment


Mean Size at Sexual Maturity $=250 \mathrm{~mm}$ TL (Thomson and Munro 1983)

APPENDIX A
Definition of Embayment Habitat Types

|  | Habitat Type | Definition of Each Habitat (after Adams and Ebersole, In Press) |
| :--- | :--- | :--- |
| Patch Reef | Isolated, high calcareous structure (not part of the contiguous reef) with <br> a vertical profile that often, but not always contains live coral cover. <br> The most important characteristic is vertical relief |  |
| Rubble | Low-relief calcareous structure composed primarily of dead/dying coral <br> fragments that are not attached to the substrate. Rubble habitat may <br> occur over extended areas or as isolated fragments within seagrass, <br> sand, or algal plain habitats |  |
| Seagrass | Monospecific or nearly monospecific, stands of Thalassia testudinum, <br> with varying densities of Syringodium filiforme mixed in. |  |
| Sand | Areas of open sand with no or very little (<10\% cover) plants or <br> coralline material represented. |  |
| Algal Plain | Sand bottom dominated by Dictyota spp., Halimeda spp., Penicillus <br> spp., Acanthophora spp., and Udotea spp., which may include sparse <br> stands of S. filiforme and T. testudinum. |  |

## Appendix B

Summary of Normality Test, One-way ANOVA, and Multiple Comparison Test Results for the Northeast Coast of St. Croix, USVI

Patch reef $=\mathrm{PR}$, algal plain $=\mathrm{AP}$, sand $=\mathrm{S}, \mathrm{R}=$ rubble, seagrass $=\mathrm{SG}$ $\mathrm{CP}=$ Cottongarden Bay, $\mathrm{TB}=$ Teague Bay, $\mathrm{YB}=$ Yellowcliff Bay

| No. | Comparison | KolmogorovSmirnov Normality Test (*1) | KruskallWallis One Way ANOVA (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
| 1 | Fish density between embayments (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 2 | Number of species density between embayments (visual census) | Failed | Not significant ( $p>0.05$ ) | Not applicable |
| 3 | Diversity index between embayments (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 4 | Evenness index between embayments (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 5 | Fish density between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, PR $>A P, P R>S, R>A P, R>S$ |
| 6 | Number of species density between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in number of species density, $\mathrm{PR}>\mathrm{AP}$, PR $>S, R>A P, R>S$ |
| 7 | Diversity index between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in diversity index, PR $>\mathrm{AP}, \mathrm{PR}>\mathrm{S}, \mathrm{PR}>\mathrm{SG}$, $P R>R$ |
| 8 | Evenness index between habitats (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 9 | O. chrysurus densities between embayments (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 10 | S. radians densities between embayments (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{CP}>\mathrm{TB}, \mathrm{CP}>\mathrm{YC}$ |

## Appendix B (continued)

Summary of Normality Test, One-way ANOVA, and Multiple Comparison Test Results for the Northeast Coast of St. Croix, USVI

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=S G$
$\mathrm{CP}=$ Cottongarden Bay, $\mathrm{TB}=$ Teague Bay, $\mathrm{YB}=$ Yellowcliff Bay.

| No. | Comparison | KolmogorovSmirnov Normality Test (*1) | KruskallWallis One Way ANOVA (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
| 11 | Haemulon unk. densities between embayments (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{YB}>\mathrm{CP}, \mathrm{TB}>\mathrm{CP}$ |
| 12 | H. bivittatus densities between embayments (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.5$ ) in fish density, <br> Y B > C P |
| 13 | Number of fish caught per trap-day between embayments (fish trap) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 14 | Number of species caught per trap-day between embayments (fish trap) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 15 | H. flavolineatum catch rates between embayments (fish trap) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 16 | O. chrysurus catch rates between embayments (fish trap) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | ( Not applicable |
| 17 | P. maculatus catch rates between embayments (fish trap) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |
| 18 | S. radians catch rates between embayments (fish trap) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in catch rates, $\mathrm{CP}>\mathrm{TB}$ |
| 19 | Number of fish caught per haul between embayments (seinehaul) | Failed | Significant <br> a <br> differences | Significant differences ( $\mathrm{p}<0.05$ ) in number of fish caught per haul, $\mathrm{TB}>\mathrm{CP}$, TB $>$ YB |
| 20 | Number of species caught per haul between embayments (seine-haul) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in number of species per haul, $\mathrm{T} \mathrm{B}>\mathrm{C} P$, $\mathrm{TB}>\mathrm{YB}$ |

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## Appendix $B$ (continued)

Summary of Normality Test, One-way ANOVA, and Multiple Comparison Test Results for the Northeast Coast of St. Croix, USVI

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=S G$

| No. | Comparison | Kolmogorov- <br> Smirnov <br> Normality <br> Test (*1) | Kruskall- <br> Wallis One <br> Way ANOVA <br> (wo | Dunn's Multiple <br> Comparison Test |
| :---: | :--- | :--- | :--- | :--- | :--- |
| 21 | C. latus catch rates <br> between embayrnents <br> (seine-haul) | Failed | Not <br> significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 22 | T. falcatus catch rates <br> between embayments <br> (seine-haul) | Failed | Not <br> significant <br> $(\gg 0.05)$ | Not applicable |
| 23 | E. jonesi catch rates <br> between embayments <br> (seine-haul) | Failed | Not <br> significant <br> $(\mathrm{P}>0.05)$ | Not applicable |

$\mathrm{CP}=$ Cottongarden Bay. TB=Teague Bav YR=Yellowcliff Bav
Notes:

1. Data must fail normality test in order to apply ANOVA on Ranks.
2. ANOVA results must be significant in order to apply Dunn's Multiple Comparison Test.

## Appendix C.

Summary of Normality Test, Two way ANOVA, and Multiple Comparison Test Results for the Northeast Coast of St. Croix, USVI.

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=S G$ $\mathrm{CP}=$ Cottongarden Bay, $\mathrm{TB}=$ Teague Bay, YB=Yellowcliff Bay.

| No. | Comparison | Kolmogorov <br> -Smirnov <br> Normality <br> Test (*1) | Two way ANOVA on Ranks (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
|  | Fish densities by Habitat and by Size Class | Failed | Significant differences ( $\mathrm{p}<0.05$ ) by treatment and between treatments | 1. Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{PR}>\mathrm{S}, \mathrm{PR}>\mathrm{SG}, \mathrm{PR}>\mathrm{AL}$, $P R>R$. <br> 2. $<5 \mathrm{~cm}$ TL fish density significantly higher ( $\mathrm{p}<0.05$ ) than all other size class densities (pooled habitats) |
| 2 | Fish densities by habitat for $<5 \mathrm{~cm}$ TL fish |  |  | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{PR}>\mathrm{S}, \mathrm{PR}>\mathrm{SG}$; |
| 3 | Fish densities by habitat for 5-10 cm TL fish |  |  | Significant differences ( $\mathrm{p}<0.05$ ) in fish density in $\mathrm{PR}>\mathrm{AP}, \mathrm{PR}>\mathrm{S}$, PR $>\mathrm{R}, \mathrm{PR}>\mathrm{SG}$ |
| 4 | Fish densities by habitat for $>10$ cm TL fish |  |  | Significant differences ( $\mathrm{p}<0.05$ ) in fish density in $\mathrm{PR}>\mathrm{AP}, \mathrm{PR}>\mathrm{S}$, PR $>\mathrm{R}, \mathrm{PR}>\mathrm{SG}$ |
| 5 | Fish densities by size class for seagrass habitat |  |  | $<5 \mathrm{~cm}$ TL fish density significantly higher ( $\mathrm{p}<0.05$ ) than densities of all other fish size classes in seagrass |
| 6 | Fish densities by size classes for patch reef habitat |  |  | $<5 \mathrm{~cm}$ TL and $5-10 \mathrm{~cm}$ TL fish densities significantly higher ( $\mathrm{p}<0.05$ ) than $>10 \mathrm{~cm}$ TL fish density in patch reefs |
| 7 | Fish densities by size classes for algal plain habitat |  |  | $<5 \mathrm{~cm}$ TL fish density significantly higher ( $\mathrm{p}<0.05$ ) than all other size classes in algal plains |
| 8 | Fish densities by size class for rubble habitat |  |  | $<5 \mathrm{~cm}$ TL fish density significantly higher ( $\mathrm{p}<0.05$ ) than all of the size classes in rubble habitat |

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## Appendix C (continued).

Summary of Normality Test, Two way ANovA, and Multiple Comparison Test Results for the Northeast Coast of St. Croix. USVI.

Patch reef $=P R$, algal plain $=A . P$, sand $=S, R=$ rubble, seagrass $=S G$ $\mathrm{CP}=$ Cottongarden Bay, $\mathrm{TB}=$ Teague Bay, $\mathrm{YB}=$ Yellowcliff Bay.

| No. | Comparison | Kolmogorov -Smirnov Normality Test (*1) | Two way ANOVA on Ranks (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
| 9 | Fish densities by size class for sand habitat |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 10 | Fish densities by embayment and by size class | Failed | No significant differences ( $\mathrm{p}>0.05$ ) by treatments or between treatments | Not applicable |
| 11 | Fish densities by embayment for $<5 \mathrm{~cm}$ TL fish |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 12 | Fish densities by embayment for 510 cm TL fish |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 13 | Fish densities by embayment for $>10 \mathrm{~cm}$ TL fish |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 14 | Fish densities by size classes for Cottongarden Bay |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 15 | Fish densities by size classes for Teague Bay |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 16 | Fish densities by size classes for Yellowcliff Bay |  |  | Not significant ( $\mathrm{p}>0.05$ ) |

1. Data must fail normality test in order to apply ANOVA on Ranks.
2. ANOVA results must be significant in order to apply Dunn's Multiple Comparison Test.

## Appendix D.

Summary of Normality Test, One-way ANOVA, and multiple comparison test Results for the Southeast Coast of St. Croix USVI.

Patch reef $=\mathrm{PR}$, algal plain $=\mathrm{AP}$, sand $=\mathrm{S}, \mathrm{R}=$ rubble, seagrass $=\mathrm{SG}$
Turner Hole=TH, Robin Bay=RB, Great Pond Bay=GP.

| No. | Comparison | KolmogorovSmirnov Normality Test (*1) | KruskallWallis One Way ANOVA (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
| 1 | Fish density between embayments (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{TH}>\mathrm{RB}, \mathrm{TH}>\mathrm{GP}$ |
| 2 | Number of species density between embayments (visual census) | I Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in number of species density, $\mathrm{TH}>\mathrm{RB}$, TH $>$ GP |
| 3 | Diversity index between embayments (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in diversity index, TH $>$ RB, TH $>$ GP |
| 4 | Evenness index between embayments (visual | Failed | Significant differences $(\mathrm{p}<0.05)$ | Significant differences ( $\mathrm{p}<0.05$ ) in evenness index, GP $>\mathrm{TH}, \mathrm{GP}>\mathrm{RB}$ |
| 5 | Fish density between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences $\{\mathrm{p}<0.05$ ) in fish density, $P R>S G, P R>S, A P>S G$, AP $>\mathrm{S}$ |
| 6 | Number of species density between habitats (visual census) | Failed | Significant differences $(\mathrm{p}<0.05)$ | Significant differences ( $\mathrm{p}<0.05$ ) in number of species density, $\mathrm{PR}>\mathrm{SG}$, PR $>S, R>S, A P>S$ |
| 7 | Diversity index between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in diversity index, |
| 8 | Evenness index between habitats (visual census) | Failed | Significant differences ( $\mathrm{p}<0.05$ ) | Significant differences ( $\mathrm{p}<0.05$ ) in evenness index, SG $>S, A P>S$ |
| 9 | Haemulon unk. densities between embayments (visual census) | Failed | Not significant ( $\mathrm{p}>0.05$ ) | Not applicable |

## Appendix D (continued).

Summary of Normality Test, One-way anova, and multiple comparison test Results for the Southeast Coast of St. Croix, USVI.

Patch reef $=$ PR, algal plain $=A P$, sand $=$ S, R. $=$ rubble, seagrass $=$ SG Turner Hole=TH, Robin Bay=RB, Great Pond Bay=GP.

| No. | Comparison | Kolmogorov- <br> Smirnov <br> Normality <br> Test (*1) | Kruskall- <br> Wallis One <br> Way ANOVA <br> $(* 2)$ | Dunn's Multiple <br> Comparison Test |
| :--- | :--- | :--- | :--- | :--- |
| 10 | S. radians densities <br> between <br> embayments (visual <br> census) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 11 | T. bifasciatum <br> densities between <br> embayments (visual <br> census) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 12 | H. bivittatus <br> densities between <br> embayments (visual <br> census) | Failed | Significant <br> differences <br> $(\mathrm{p}<0.05)$ | Significant differences <br> $(\mathrm{p}<0.05)$ in fish density, <br> TH $>$ RB |
| 13 | Number of fish <br> caught per trap-day <br> between <br> embayments (fish <br> trap) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 14 | Number of species <br> caught per trap-day <br> between <br> embayments (fish <br> trap) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 15 | Number of fish <br> caught per haul <br> (seine-haul) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |
| 16 | Number of species <br> caught per haul <br> between <br> embayments (seine- <br> haul) | Failed | Not significant <br> $(\mathrm{p}>0.05)$ | Not applicable |

* Notes:

1. Data must fail normality test in order to apply ANOVA on Ranks.
2. ANOVA results must be significant in order to apply Dunn's Multiple Comparison Test.

## Appendix E

Summary of Normality Test, Two-way .ANOVA, and Multiple Comparison Test Results for the Southeast Coast of St. Croix, USVI.

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=$ SG Turner
Hole $=$ TH, Robin Bay=RB, Great Pond Bay=GP.

| No. | Comparison | KolmogorovSmirnov Normality Test (*1) | Two Way ANOVA on Ranks (*2) | Dunn's Multiple Comparison Test |
| :---: | :---: | :---: | :---: | :---: |
| 1 | Fish densities by habitat and by size class | Failed | Significant differences ( $\mathrm{p}<0.05$ ) by treatment and between treatments | 1. Significant differences $(\mathrm{p}<0.05)$ in fish densities, PR $>\mathrm{R}$ <br> 2. $<5 \mathrm{~cm}$ TL fish densities significantly higher ( $\mathrm{p}<0.05$ ) than all other size class densities (habitats pooled) |
| 2 | Fish density by habitat for $<5 \mathrm{~cm}$ TL fish |  |  | Significant differences ( $\mathrm{p}<0.05$ ) in fish density, $\mathrm{PR}>\mathrm{S}, \mathrm{PR}>\mathrm{SG}, \mathrm{AP}>\mathrm{S}$, AP>SG |
| 3 | Fish density by habitat for 540 cm TL fish |  |  | Significant differences, ( $\mathrm{p}<0.05$ ) in fish densities, $P R>S G, \mathrm{PR}>\mathrm{S}, P R>A P$ |
| 4 | Fish density by habitat for $>10 \mathrm{~cm}$ TL fish |  |  | $\begin{aligned} & \text { (Significant differences, } \\ & (\mathrm{p}<0.05) \text { in fish densities, } \\ & \mathrm{PR}>\mathrm{S}, \mathrm{PR}>\mathrm{R}, \mathrm{PR}>\mathrm{SG} \end{aligned}$ |
| 5 | Fish density by size class for seagrass habitat |  |  | Not significant ( $\mathrm{p}>0.05$ ) |
| 6 | Fish density by size class for patch reef habitat |  |  | $<5 \mathrm{~cm} \mathrm{TL}$ and 5-10 cm TL fish densities significantly higher ( $\mathrm{p}<0.05$ ) than $>10 \mathrm{~cm} \mathrm{TL}$ fish density in patch reefs |
| 7 | Fish densities by size classes for algal plain habitat |  |  | $<5 \mathrm{~cm}$ TL fish density significantly higher ( $p<0.05$ ) than all size classes in algal plains |

## Appendix E (continued)

Summary of Normality Test, Two-way ANOVA, and Multiple Comparison Test Results for the Southeast Coast of St. Croix, USVI.

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=$ SG Turner
Hole $=$ TH, Robin Bay $=$ RB, Great Pond Bay=GP.
$\left.\begin{array}{|l|l|l|l|l|}\hline \text { No. } & \text { Comparison } & \begin{array}{l}\text { Kolmogorov- } \\ \text { Smirnov } \\ \text { Normality } \\ \text { Test (*1) }\end{array} & \begin{array}{l}\text { Two Way } \\ \text { ANOVA on } \\ \text { Ranks (*2) }\end{array} & \begin{array}{l}\text { Dunn's Multiple } \\ \text { Comparison Test }\end{array} \\ \hline 8 & \begin{array}{l}\text { Fish densities by } \\ \text { size class for rubble } \\ \text { habitat }\end{array} & & & \begin{array}{l}<5 \mathrm{~cm} \text { TL and 5-10 cm TL } \\ \text { fish densities significantly } \\ \text { higher (P }<0.05) ~ t h a n ~\end{array} 10 \mathrm{~cm} \\ \text { j }\end{array}\right\}$

## Appendix E (continued)

Summary of Normality Test, Two-way ANOVA, and Multiple Comparison Test Results for the Southeast Coast of St. Croix, USVI.

Patch reef $=P R$, algal plain $=A P$, sand $=S, R=$ rubble, seagrass $=$ SG Turner
Hole $=\mathrm{TH}$, Robin $\mathrm{Bay}=\mathrm{RB}$, Great Pond $\mathrm{Bav}=\mathrm{GP}$,

| No. | Comparison | Kolmogorov- <br> Smirnov <br> Normality <br> Test (*1) | Two Way <br> ANOVA on <br> Ranks (*2) | Dunn's Multiple <br> Comparison Test |
| :--- | :--- | :--- | :--- | :--- |
| 16 | Fish densities by <br> size classes for <br> Great Pond |  |  | $<5 \mathrm{~cm}$ TL and 5-10 cm IL <br> fish densities significantly <br> higher (p $<0.05)$ than $>10 \mathrm{~cm}$ <br> TL fish density in Great Pond <br> Bay |

* Notes:

1. Data must fail normality test in order to apply ANOVA on Ranks.
2. ANOVA results must be significant in order to apply Dunn's Multiple Comparison Test.

[^0]:    *not caught

[^1]:    *note: Total area of habitat surveyed in transects.

[^2]:    Legend: 冨Cottongarden Bay Teague Bay $\uparrow$ Yellowivliff Bay

